

THE ELK OF GRAND TETON AND
SOUTHERN YELLOWSTONE NATIONAL PARKS

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" . . . individually we contribute little or nothing to the truth, by the union of all a considerable amount is amassed."

Aristotle (384-322 B.C.)

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ABSTRACT

A 5-year field study was conducted in and adjacent to Grand Teton and southern Yellowstone National Parks. Objectives were to obtain reference information on the habits, population dynamics, and ecology of elk; and evaluate management programs. The two parks provided spring, summer, and fall habitat for the greater portion of the studied elk population which wintered on the National Elk Refuge immediately south of Grand Teton. The study area extended about 55 miles northward from the refuge's south boundary into the mountain and plateau regions of southern Yellowstone Park. Vegetation on the area was classified. Successional relationships were diagramed. The area's history from a summer hunting ground for Indians, through fur trade, settlement, and agricultural eras to the development of tourism was reviewed.

The elk wintering on the refuge were part of the Jackson Hole herd which historically wintered within and adjacent to extensive valley lands north and south of the refuge and the adjoining Gros Ventre and Buffalo river drainages. Early records suggest this herd contained up to 20,000 animals and was distinctly separate from elk which wintered in the Green River drainage. Migration and distribution studies indicated that about 29 percent of the elk wintering on the refuge summered in Grand Teton; 42 percent, in southern Yellowstone Park. Most of the remainder summered over extensive national forest lands adjoining the

two parks. A resident summer herd of over 1,000 animals became established within Grand Teton valley areas that had been closed to hunting for about 10 years. Examinations of reproductive organs suggested that about 89 percent of the female elk older than yearlings became pregnant. Mortality of newborn calves probably averaged about 60 percent. Differences between reflected annual increase rates by calf proportions in winter herds, their realized recruitment as yearlings, and "best fit" comparisons of calculated herd sizes with periodic censuses suggested that a combined 2 to 3 percent mortality of calves and adults occurred during and after most winters of average severity; at least 15 percent during and after more severe winters. Overwinter mortality that averaged about 5 percent over mild, average, and severe winters was probably partially density-independent within the range of population size accommodated by variable environmental conditions.

Census records for 57 years since 1911 showed relatively "stable" population trends for the herd wintering on the refuge. Winter numbers fluctuated within a 6,000-8,000 range about 65 percent of this time, a 5,000-9,000 range 78 percent of this time, and a 5,000-10,000 range with a 98 percent frequency. Fluctuations showed no obvious relationship to hunting removals until after 1950. Elk numbers for the Jackson Hole herd as a whole declined about 30 percent over the 57-year period. Declines mainly occurred in herd segments that had substantial portions of their winter range appropriated for human

settlement and agriculture, or in groups that wintered off the refuge on smaller state feedgrounds established to reduce elk conflicts with agriculture. Overall declines coincided with an approximate loss of one-third of the herd's original winter range and substitutions of livestock for elk grazing on other wintering areas.

Dispersals to spring ranges on Grand Teton usually occurred during the first half of May. The use of different areas for calving and rates of migration to Yellowstone summer ranges were variably influenced by snow accumulations in mountain passes. Progressive movements from low to high elevations on summer ranges usually occurred to about late July. Differences occurred with sex and age, parous and nonparous females, and molesting insect activity. Reverse movements usually started in August. Marked animals showed elk from widely separate winter herds were intermingled on Yellowstone summer ranges. Interchanges of animals between winter herds were limited. Records from fall migration studies showed that elk groups that summered in and/or migrated through more accessible hunting areas outside Grand Teton boundaries progressively declined through the 1950's. Increases occurred in groups summering in national parks or migrating through less accessible hunting areas outside parks.

Relative use of different habitats was shown by observation samples of 82,223 elk in valley areas and 20,017 in mountain areas.

Use of bunchgrass-shrub, sagebrush, valley meadow, forest, and agricultural types averaged 30, 29, 21, 17, and 3 percent, respectively, for April through December periods in valley areas. Use of forest, herbland, forest park burn, valley meadow, and subalpine meadow types averaged 37, 30, 16, 9, 6, and 2 percent, respectively, for June through October periods in mountain areas. Differences occurred with seasonal periods and years.

Food habits information was obtained from 262,602 instances of recorded plant use at 473 elk feeding sites. An average yearlong food habit of about 51 percent grass and grasslike plants, 26 percent forbs, and 23 percent shrubs was calculated for free-ranging elk. Differences in forage class and plant species use showed elk were extremely versatile and generalized feeders with a capacity to contend with a wide spectrum of environmental change. Vegetation measurements showed elk maintained biotic disclimax on limited ridgetop and upper slope sites that were kept free of snow by wind or first thaws. Variable snow conditions and the foraging actions of the elk themselves appeared to prevent the animals from progressively depleting their main food sources on bottomland, upland swale and slope areas that were usually snow covered from November through March. The animals hastened the replacement of seral willow or aspen when stands reached late successional stages or remnant status. Elk effects which represented a departure from natural

relationships appeared to be limited to wintering areas within 1 mile of feedgrounds. Measurements of elk and other biotic effects on spring and high elevation summer ranges suggested these were limited or of a temporary nature and did not represent departures from natural relationships.

Accumulated information from elk distribution and migration studies led to specific objectives for cooperative State and Park Service management programs after 1963. Long-term objectives were to restore historical distributions and migrations that had been altered by unequal hunting removals and reduce the need to hunt elk in Grand Teton Park. Specific short-term objectives were to halt October migrations of Grand Teton summer herds to refuge winter ranges, reduce late migrating Yellowstone elk groups that had increased during the 1950's, and allow compensating increases in other groups that summered and/or migrated through areas outside park boundaries. October migrations were largely halted after 1964. Management programs that involved the use of split hunting periods, special permit areas, and kill quotas for different herd groups appeared to start trends toward accomplishing distribution and/or migration changes after 1963.

A section on ecology integrates study findings into an overall account of suggested relationships within and between the elk population and its environment. Climate, winter weather and food, plant

succession, predators, scavengers, parasites, disease, and man are related to the elk as environmental influences. Artificial feeding was suggested to have population consequences to the extent that it did not provide net gains in energy over what calves and pregnant females would obtain by freeranging. Considerations of elk behavior indicated that the animals influenced their own distributions in the absence of overriding environmental influences. Social disorder on feedgrounds and conservation of energy relationships are discussed. Interpretations of elk habitat relationships required considerations of natural selection and plant successional processes, the ecological completeness of winter habitats, and distinctions between food sources that did or did not have population consequences. Intraspecific competition for food and environmental influences from winter weather, predators, scavengers, and disease were considered to represent the natural regulatory complex on presettlement elk populations. The regulation of present populations differed to the extent that man increased or decreased the intensity of natural regulatory influences and substituted himself for the original predator-scavenger fauna. Density-independent mortality precluded maintaining highly stable elk numbers and fully substituting hunting removals or artificial feeding for all severe weather influences.

Man's hunting appeared more efficient than original predator-scavenger complexes in reducing extreme fluctuations in elk numbers.

It was less efficient in allowing the elk population to maintain its numbers and distributions in relation to suitable habitats and available food sources. Purposes of the two national parks and other cooperating agencies are discussed relative to the studied elk population. Cooperative management programs to restore historical distributions and migrations are expected to progressively reduce the need for large scale hunting within Grand Teton and obtain desired elk kills by recreational hunting outside park boundaries.

INTRODUCTION

The Rocky Mountain elk (Cervus canadensis nelsoni) is part of the native fauna of Grand Teton and Yellowstone National Parks. The ancestors of the species are generally believed to have crossed the Bering Sea land bridge to North America during periods of Pleistocene glaciation (Frick, 1937). Fossil records show the presence of elk in Alaska at least 100,000 years ago (Péwé and Hopkins, 1965). An ice-free passage to central North America may not have existed until an interglacial period between 35,000 and 25,000 years ago and not again until after 14,000 years ago (Hopkins, 1965). It seems probable that a close ancestor to the Rocky Mountain elk arrived in central North America between 35,000 and 25,000 years ago. The present modern species probably moved from and back into regions that became Grand Teton and Yellowstone Parks with the advance and retreat of less extensive intermountain glaciers which reached their maximum 12,000 to 10,000 years ago.

The lands within the man-made boundaries of Grand Teton and southern Yellowstone Park did not represent a complete ecological unit for the greater portion of the elk population which is the subject of this bulletin. The majority of the animals used national park lands during late spring, summer, and early fall. They migrated over or foraged on lands outside park boundaries during other seasonal periods.

Outside the two parks, the elk or the lands they used were subject to the jurisdictions of the State of Wyoming, U. S. Forest Service, or Refuge Branch of the Bureau of Sport Fisheries and Wildlife. These agencies had objectives directed toward the management of the elk or its habitat to provide recreational hunting. Grand Teton and Yellowstone objectives were encompassed in their primary purpose of preserving natural environments and native plant and animal life for their scenic, educational, cultural, or scientific values.

The situation where agencies with different responsibilities or objectives in public service were concerned with the same elk population made interagency cooperation necessary. Cooperative management of the elk on portions of Grand Teton Park was provided for by a special law. Wyoming Game and Fish Commission personnel reviewed their proposed elk management programs for lands outside parks with other agencies. Field studies relating to the elk were coordinated through a committee of administrators and technical personnel from each agency.

This bulletin presents the results of field studies by the author from June 1962 through May 1967. Other pertinent information from a variety of sources is included for comparison purposes or to maintain the continuity of records. Study objectives were to obtain basic reference information on elk habits, population dynamics, and ecology and to evaluate the effectiveness of management programs within Grand Teton Park.

THE STUDY AREA

The study area includes most of the land used by Grand Teton and southern Yellowstone elk (Figure 1). It extends about 55 miles north of the south boundary of the National Elk Refuge near Jackson, Wyoming, through Grand Teton Park and Teton National Forest lands into Yellowstone Park. Most of the area is on the west side of the Continental Divide in the headwaters system of the Snake River. A comprehensive account of the area's geology is given by Fryxell (1930), Love and Reed (1968). Physiographic features and soils show the effects of glacial action. Glaciers extended onto Grand Teton valley floor as recently as 9,000 years ago.

Physiography

The main physiographic features on the study area and adjoining lands are shown on Figure 2. The map regions shown as Jackson Hole, the Mount Leidy and Pinyon Peak highlands, the Snake River drainage slopes of the Teton Mountain, Gros Ventre, and Hoback Ranges, and the east slope of the Snake River Range are called "the Jackson Hole area." The portion of the valley south of the Snake River outlet from Jackson Lake is called "the Jackson Hole valley."

A comparatively flat glacial outwash plain forms the floor of the Jackson Hole valley. Other features on the plain are glaciated

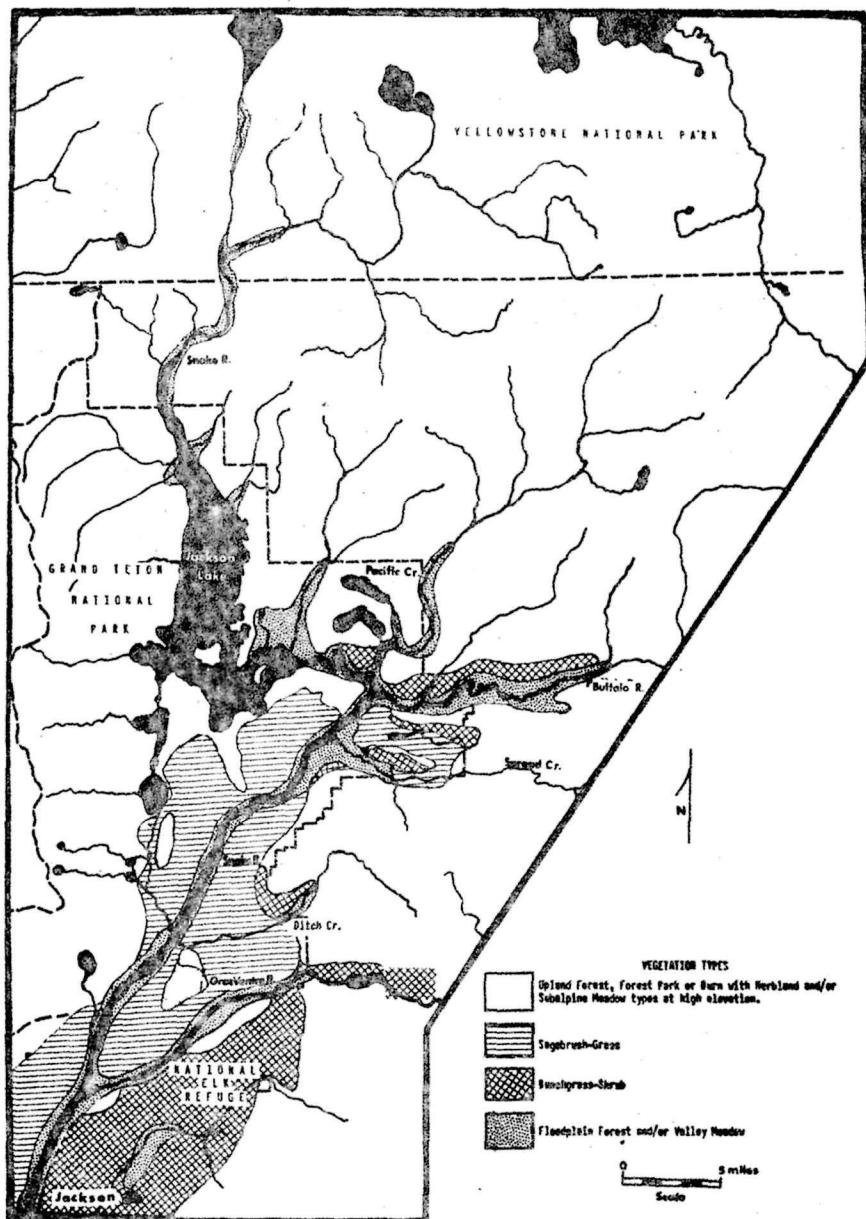


Fig. 1. Map showing major vegetation types on the study area.



Fig. 2. Physiographic map of the study area and adjoining lands (by permission of Dr. S. H. Knight from 1956 Jackson Hole Field Conference Guidebook, Wyoming Geological Assn., p. 7).

ridges that project through the surrounding outwash deposits, terrace benches along the sides of past and present water courses, potholes formed from melting ice blocks, alluvial fans, and braided stream bottoms. Foothill ridges and slopes border the valley on the east. Glacial moraine deposits border the valley on the north and west and extend as projections onto the outwash plain.

A series of high ridges, mountains, and plateaus occur from the north end of the valley into southern Yellowstone Park. Elevations range from a minimum of about 6,200 feet on the refuge to maximums of over 13,700 feet on the Teton Range, and slightly above 10,000 feet for the highest mountain and plateau tops within southern Yellowstone Park.

The outwash plain soils on the valley floor are poorly developed and contain a high proportion of rounded cobblestones. Glacial moraine and alluvial fan soils contain relatively greater proportions of fine textured material in variable mixtures with angular rock and cobble. Ridge and foothill slopes have mostly fine textured residual soils, but coarse textures occur where resistant rock formations are exposed or overlying glacial deposits are present. Soils along water courses range from layer cobble with gravel and sand in interspaces to developed sand, silt, and clay loams. Mountain and plateau soils are extremely variable, ranging from deep residual loams to very coarse textures developed from rock talus or volcanic rhyolite.

Climate

The area has short, cool summers and cold winters. U. S. Department of Commerce records for Moran at the northern limits of the Jackson Hole valley show a 30-year (1936-1965) mean annual temperature of 34.8° F. (Houston, 1967). The mean temperature for July as the warmest month is 60° F. January as the coldest, 10.3° F. About three-fourths of annual precipitation of 22.2 inches for the period fell as snow.

West to east and north to south gradients in lessened snow depths were apparent. Midwinter snow depths on the west side and north of the Jackson Hole valley commonly ranged from 6 to 3 feet, as compared to 3 to less than 1 foot depths on the east side or in the southern portions of the valley.

Moran was at the northern limit of the area used by any appreciable numbers of wintering elk. December through March snow measurements at this station were used to reflect the relative magnitude of this environmental influence between 1961 and 1967 (Table 1).

The period from November 16 through April 15 was considered the winter season. April 16 through June 15 in valley areas, and May 16 through June 30 in the mountain areas, was spring. June 15 or 30 to September 15 was summer. Fall extended from September 16 through November 15.

Table 1.--Mean monthly snow depths for December through March periods, 1961-1967.

Years	December-March mean monthly snow depths (inches)	Relative snow depths ¹
1961-62	40	High
1962-63	19	Low
1963-64	34	Average
1964-65	43	High
1965-66	30	Average
1966-67	32	Average
1951-1967 Mean	32	

¹ ±4 inches from the mean was considered average.

Vegetation

Scientific and common names for plants mentioned in this bulletin are listed in Appendix I. These generally follow Davis (1952) or Booth and Wright (1959). Vegetation was classified during field inspections and by analyzing measurements of taxa in representative stands. Measurements were by the author (Appendix II-IV) and other workers on the area (Martinka, 1965; Oswald, 1966; Houston, 1967), following Daubenmire (1959), Parker (1951), and Cottam and Curtis (1949).

Figure 1 shows the major vegetation types on the area. Table 2 lists information on their general distribution and the characteristic as well as important associated plants for the more representative stands within each type. Various mixtures of characteristic and important associated plants occurred in transitional or intermediate successional stages.

Characteristic plants occurred consistently and usually made up the greater portion of the plant crown cover in representative or successional related stands. Important associated plants for other than forest stands had a crown cover that was slightly less, equal to, or greater than characteristic plants. Important associated plants under forest canopies made up one-fourth or more of the ground level crown cover in distinct stands, or had a consistent high frequency of occurrence in successional related stands.

Table 2.--Classification of vegetation on study area.

Vegetation Type	General Distribution	Characteristic Plants	Associated Plants
Bunchgrass-Shrub	Residual and glacial till soils on ridgetops and south, east, and west slopes in valley areas	Bluebunch Wheatgrass-Douglas Rabbitbrush Climax Big Sagebrush-Bitterbrush Stage Needle and Thread-Sandberg Bluegrass Disclimax	Ricegrass Rubber Rabbitbrush Serviceberry Threetip Sagebrush Idaho Fescue
Sagebrush	Glacial outwash plain, stream-cut terraces and alluvial fans	Big Sagebrush Stage Big Sagebrush-Bitterbrush Stage	Bluebunch Wheatgrass Idaho Fescue Balsamroot Lupine Buckwheat Low Sagebrush
Valley Meadow	Alluvial soils along water courses below 7500 feet.	Willow-Sedge Stage Tufted Hairgrass-Sedge Stage Sedge-Bluegrass Stage	Rush Shrubby Cinquefoil Silver Sagebrush Dandelion
Floodplain Forest	Stream alluvium up to about 7000 feet.	Narrowleaf Cottonwood Stage Blue Spruce Climax	Russet Buffaloberry Red Dogwood Willow Thinleaf Alder Balsam Poplar
Upland Forest	Glacial till, north and west slopes within the valley and on a variety of substrates and exposures in mountain areas up to 10,000 feet.	Lodgepole Pine Stage Aspen Stage Douglas-Fir Climax Whitebark Pine Climax Engelmann Spruce-Subalpine Fir Climax	Geyer's Sedge Pinegrass Arnica Large Huckleberry Dwarf Huckleberry Pachystima Menziesia Spirea
Burn	Upland forest stands burned since 1930.	Geyer's Sedge-Lodgepole Pine Disclimax Geyer's Sedge-Subalpine Fir Disclimax	(as in Upland Forest)

Exceptions to this classification system were the herbland type, where the crown cover of various forbs was usually greater than the two characterizing grasses, and the bunchgrass-shrub type, where big sagebrush and/or other shrubs had a greater crown cover than bluebunch wheatgrass on rocky or more mesic sites. Plant species that characterized successional advanced stands which appeared capable of persisting under the present climate are given a "climax" suffix. Plants characterizing seral vegetation that could ultimately be replaced by climax stands are called a "stage." Plants characterizing stands modified by recent fires or animal use are called a "disclimax." The classification system represents a modification of criteria used by Weaver and Clements (1938) and Daubenmire (1953).

Figures 3 and 4 show what appeared to be the general successional patterns for vegetation in valley and mountain areas. The four types of climax vegetation below 8,500 feet appeared to develop from a variety of seral stands on different sites, exposures, and soils. Successional patterns for subalpine vegetation above 8,500 feet appeared to be comparatively simple. The dual role of mountain brome, slender wheatgrass, and tufted hairgrass in climax stands at high elevations and seral stands at low elevations was evident. Some of the associated forbs for the herbland type probably had similar roles. Dual successional roles for tree species have been reported by Daubenmire (1953).

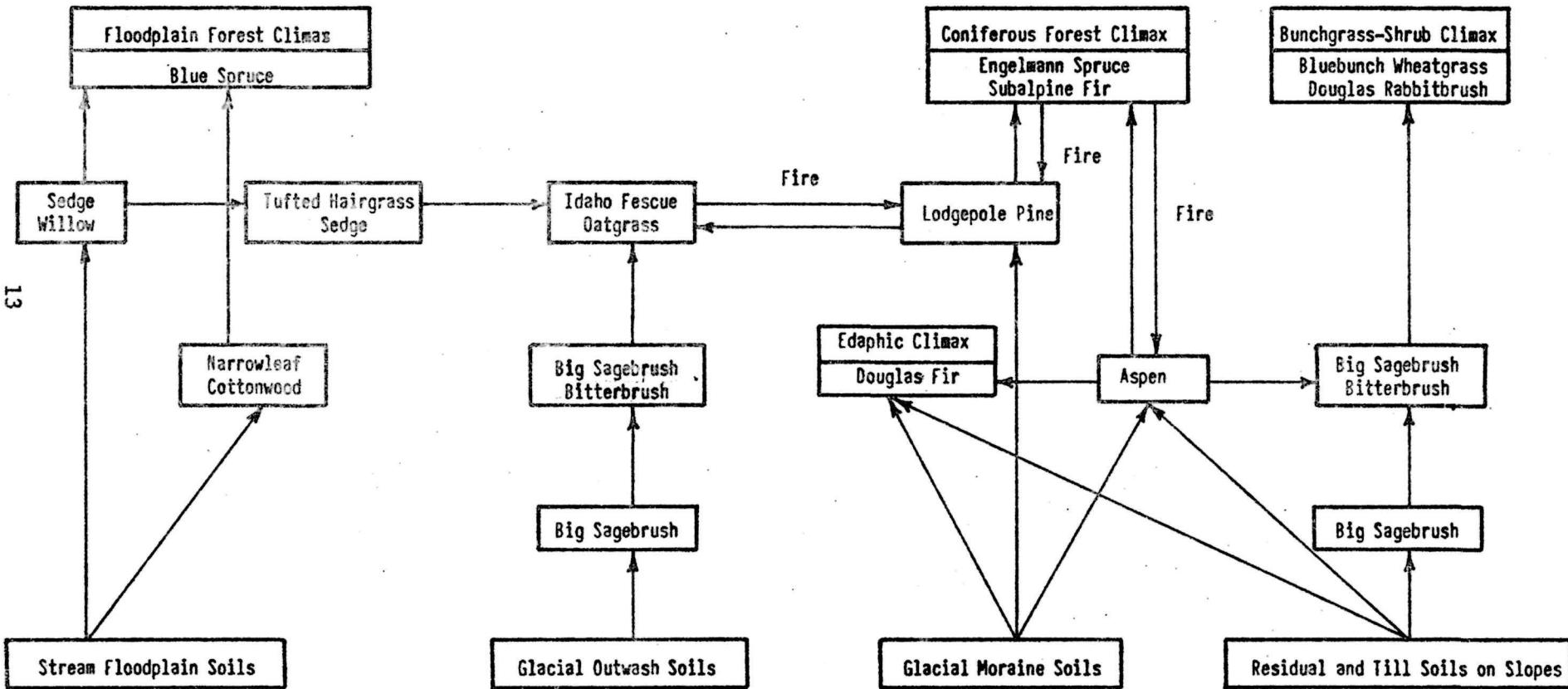


Fig. 3. Plant succession relationships at elevations from 6200-7500 ft. within valley areas.

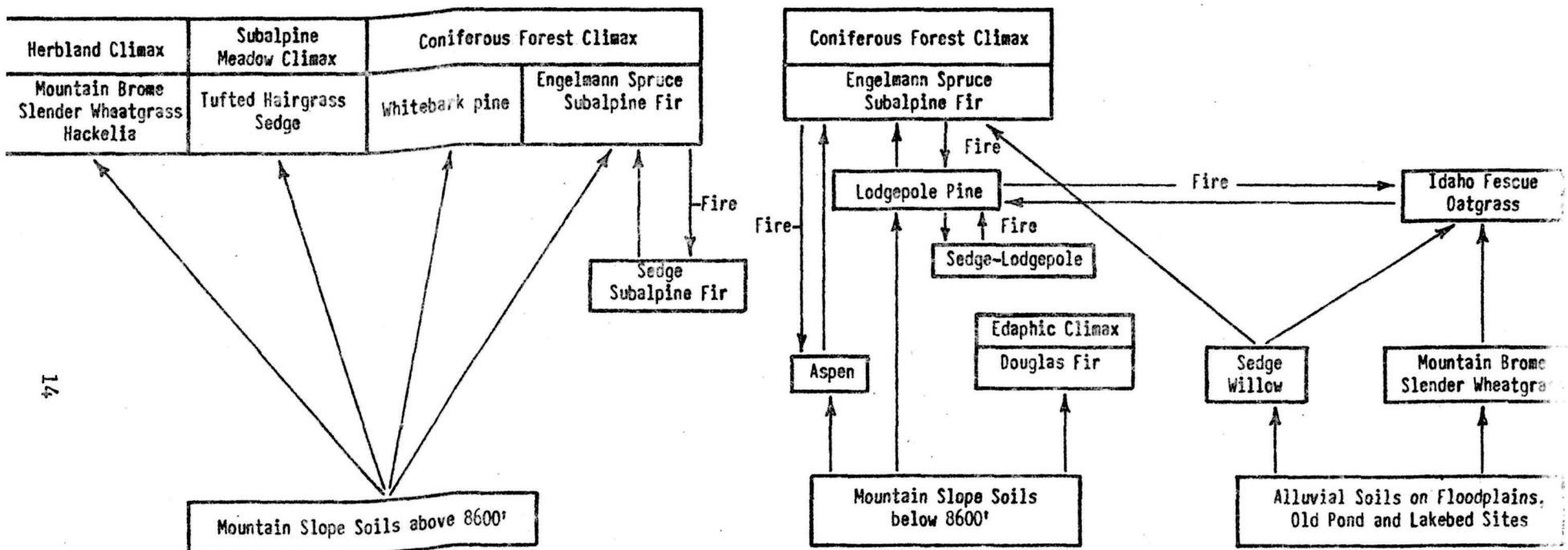


Fig. 4. Plant succession relationships at elevations from about 7000-10,000 ft. within northern mountain areas.

Most Douglas-fir and whitebark pine stands on the study area were considered edaphic climaxes, but on some more mesic sites they were probably seral to Engelmann spruce and subalpine fir. Ellison (1954) concluded that scattered stands of Engelmann spruce and subalpine fir on rocky outcrops in the herbland type were seral to herbland. This was on the Wasatch Plateau in Utah. In this study such stands occurred on rocky sites in both the herbland and subalpine meadow types and frequently contained whitebark pine. They were considered to be edaphic climaxes rather than a seral stage. All lodgepole pine stands were considered seral, but it is recognized that some stands could be considered "fire climax" as proposed by Daubenmire (1968). Fire and the biotic influences of wild ungulates in antler rubbing and foraging contributed to maintaining the forest park type.

The classifications of vegetation and judgments of successional patterns made during this study should be considered tentative. Revisions and refinements from more detailed studies are expected. The general appearance of the major vegetation types is shown on Figure 5.

History

Yellowstone National Park was established in 1872. The main east face of the Teton Mountain Range became Grand Teton National Park in 1929. Northern Jackson Hole valley lands and adjoining mountain

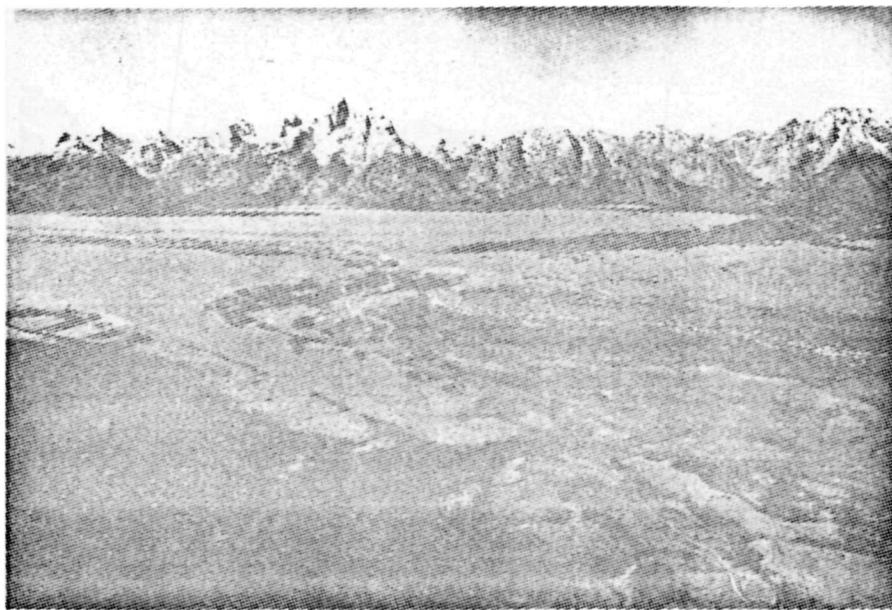


Fig. 5a. Approximate center of Grand Teton Park showing extensive sagebrush type on the outwash plain, the valley meadow, and bottomland forest types along the Snake River flood plain, and the upland forest type on glacial moraine sites.

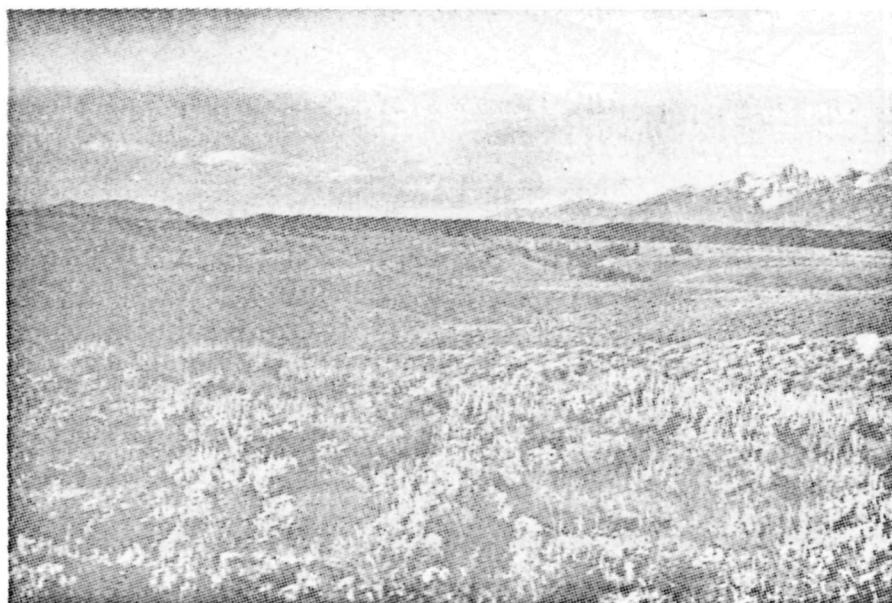


Fig. 5b. The sagebrush type in a pothole area.

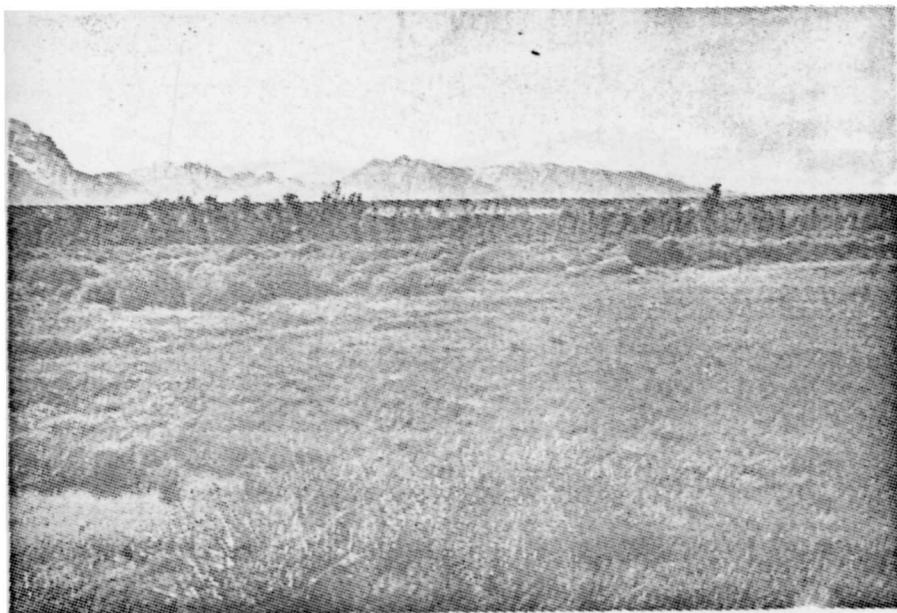


Fig. 5c. Valley meadow and bottomland forest types on Snake River flood plain.

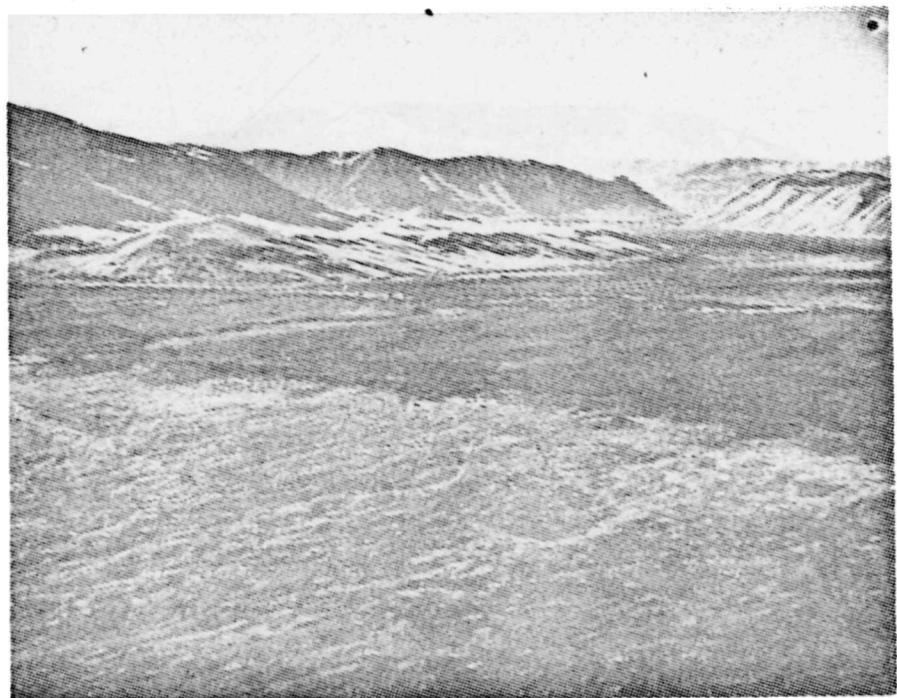


Fig. 5d. Valley meadow type on National Elk Refuge bottomlands and the bunchgrass-shrub type on adjoining slopes.



Fig. 5e. Bunchgrass-shrub type on south slopes with bottomland developed as a hayfield.

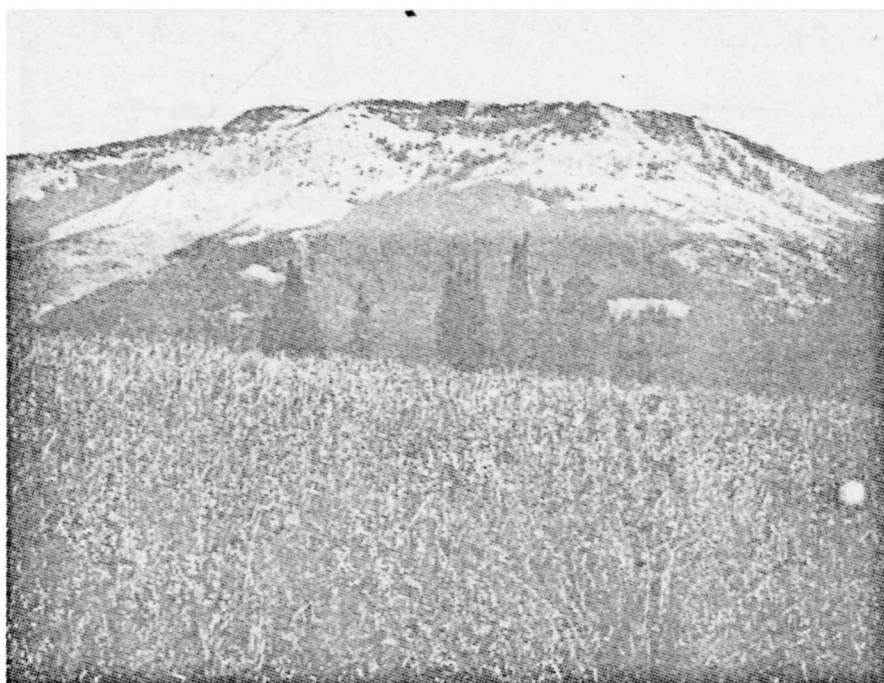


Fig. 5f. Herbland type on upper portions of Chicken Ridge and in the foreground.

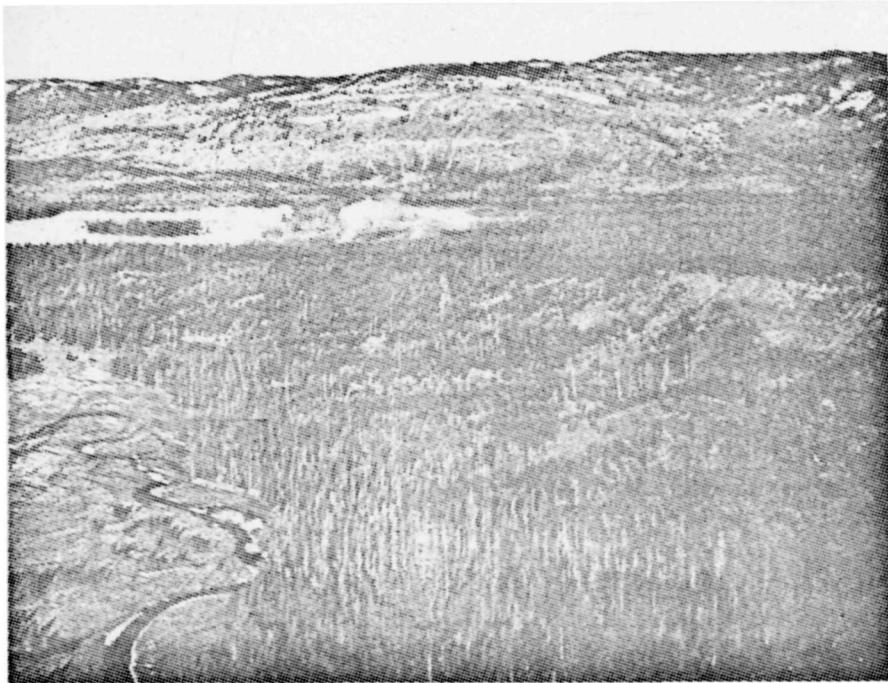


Figure 5g. Interspersed forest park, burn, and upland forest types with herbland above the upper half of Red Creek Ridge.

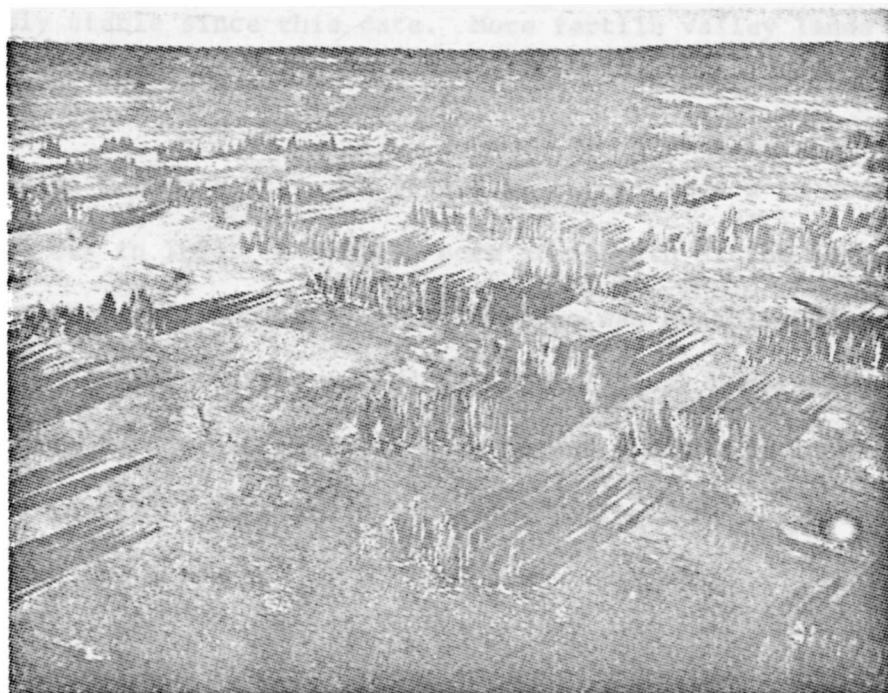


Fig. 5h. Extensive subalpine meadow type with scattered Engelmann spruce-subalpine fir stands on Pitchstone Plateau.

areas were given national monument status in 1943. These were incorporated into the present boundaries of Grand Teton in 1950. The Forest Service assumed jurisdiction over its lands in 1897.

The following summary of Jackson Hole history is largely from Hayden (1956). The region was originally a summer hunting ground for Bannock, Shoshone, and Crow Indians. Organized fur trapping occurred from about 1811 to 1840. After the collapse of the fur trade only occasional expedition parties, trappers, gold prospectors, and hunters visited the region. First permanent settlement occurred in 1884 and the human population reached 65 persons in 1889. Settlers increased and cattle raising became the chief industry. Cattle numbers increased from about 3,000 head in 1910 to about 15,000 by 1935 and have remained relatively stable since this date. More fertile valley lands were developed as hayfields.

Wyoming became a state in 1890. It established a 10-month closed season on elk in 1895. A 570,000-acre area north of the Jackson Hole valley to the south boundary of Yellowstone was established as a game preserve in 1905. A brief period of illegal "tusk hunting" prompted local vigilante action. A hunting guide and outfitting industry was operative in 1895. This and a later developing dude ranch industry continue to the present.

Publicized elk dieoffs during severe winters and conflicts between elk and livestock-raising operations led to the establishment

of the National Elk Refuge in 1913. The refuge did not reach its present size of about 22,700 acres until 1950. The Forest Service started to exclude cattle and horse grazing from important elk wintering areas on public lands in 1919. Many of the smaller ranchers unsuccessfully petitioned Congress to make the Jackson Hole valley a recreational area in 1925. The Wyoming Legislature passed a game damage law in 1939 which allowed claims for elk damage.

Yellowstone Superintendents reported that highly organized market hunting, poaching, and extremely liberal legal hunting occurred inside the park before 1894. The more remote southern portions of the park, like the Jackson Hole area, probably escaped this exploitive era. The Lacey Act of 1894 established laws and fines to protect Yellowstone wildlife. Comparatively few tourists visited the parks until after World War II. Visits greatly increased after 1946 and tourism became the major industry in the general region bordering Grand Teton and Yellowstone Parks (Rajender et al., 1967)

THE ELK POPULATION

The elk that winter on lands within the Jackson Hole, Buffalo River, and Gros Ventre River valleys are collectively called the Jackson Hole herd. A division of this herd into its north and south segments, major winter herds, summer segments and groups is shown in Table 3. The elk that summer in Grand Teton and southern Yellowstone Parks are shown to be part of the refuge winter herd as well as the northern herd.

The period before the winter of 1955-56 is considered the past; since this date, the present. Comparisons of past and present numbers that start from the fur trade period could only be made by using records for the entire Jackson Hole herd. Records of elk numbers on and adjacent to the area that became the National Elk Refuge start from the winter of 1911-12.

Jackson Hole Herd

Numbers

Records before 1900 are limited to general narrative accounts. Osborne Russell kept a detailed journal on his trips through the Jackson Hole and Yellowstone areas during 1834 and 1843 (Haines, 1965). In the Jackson Hole area during July he made such comments as: "This valley, like all other parts of the country, abounds with game." and "Game is plenty and the river and lake abounds with fish." He specifically

Table 3.--Division of Jackson Hole Elk Herd into major population segments, winter herds, summer segments and groups

Major population segments	Winter herds	Summer herd segments	Summer groups	Migratory segments
NORTHERN HERD	Refuge	Refuge	Valley	Nonmigratory
		Grand Teton Park	Valley Mountain	Grand Teton
		Southern Yellowstone Park	Pitchstone Plateau Central Mountain Two Ocean Plateau	
			Glade Creek Pilgrim Creek Pacific Creek Ditch Creek Sheep Mountain	North
		Teton National Forest	Two Ocean Plateau Buffalo Fork Gros Ventre	
	Gros Ventre			Gros Ventre
			Hoback Lower Snake	
SOUTHERN HERD	South Park-Hoback			South

reported that the valley on the west side of the Teton Range " . . . abounds with Buffalo Elk Deer antelope etc. . . ." Reports of killing these animals for food were made by Russell and others traveling through the Jackson Hole area. In the Yellowstone Lake area during August, Russell wrote " . . . we found the whole country swarming with Elk . . ." Other general summer observations of abundant elk and other wildlife in southern Yellowstone from 1870 through 1876 expedition reports have been reviewed by Murie (1940). These strongly refudiate opinions that elk and other wildlife were originally scarce in the mountains. Sheldon (1927) cited an 1887 account of a Jackson Hole trapper reporting that 15,000 elk wintered in the valleys of the Shoshone and Snake. This should probably read Shoshone or Snake. Evermann (1892) reports that early maps labeled the Snake River as the Shoshone. A presettlement account of elk wintering along the Snake River bottoms in the 1870's with greater numbers near the south end of the Jackson Hole valley is related by Murie (1951).

Records of elk numbers between 1900 and 1910 are mainly from estimates by Nowlin (1904 and 1909) and Preble (1911). These varied from 20,000 to 25,000 animals. Kneipp (1915) used a partial ground count of over 13,500 animals to estimate that at least 17,000 elk were present during the winter of 1911-12.

Records of numbers in winter herds from 1915-16 on are from periodic ground and/or aerial counts (after 1927). Ground counts

were made by crews of Federal agency and State game commission personnel. A tabulation of counts obtained over the 40-year period up to the 1955-56 winter is presented by Anderson (1958). An average winter count of about 20,000 elk (19,238 to 22,035) was obtained during five winters within the first 20-year period from 1915 to 1935. An average of 16,300 elk per winter (15,014 to 17,902) was counted during six winters within the next 19-year period. An average winter count of about 14,000 elk (11,057 to 17,924) was obtained during six of the winters since 1955-56 (Yorgason, writ. comm., 1968). The counts since 1955 did not always include scattered groups off main wintering areas and the actual average may be slightly higher than shown.

Winter Distributions

A map of Jackson Hole valley wintering areas (Figure 6) shows elk originally wintered on bottomlands and slopes in the Buffalo River valley along the east side of the Jackson Hole valley and south from Ditch Creek through the valley floor into the adjoining Hoback drainage. Increasing human settlement, agriculture, hunting, and the development of elk feed grounds progressively changed winter elk distributions after 1910. The proportion of the herd wintering on established feed grounds increased to about 48 percent by 1935 and 86 percent by 1956 (Anderson, 1958).

Small bands of elk still use historical winter ranges north of Ditch Creek (Figure 6). South of this creek the greater portion

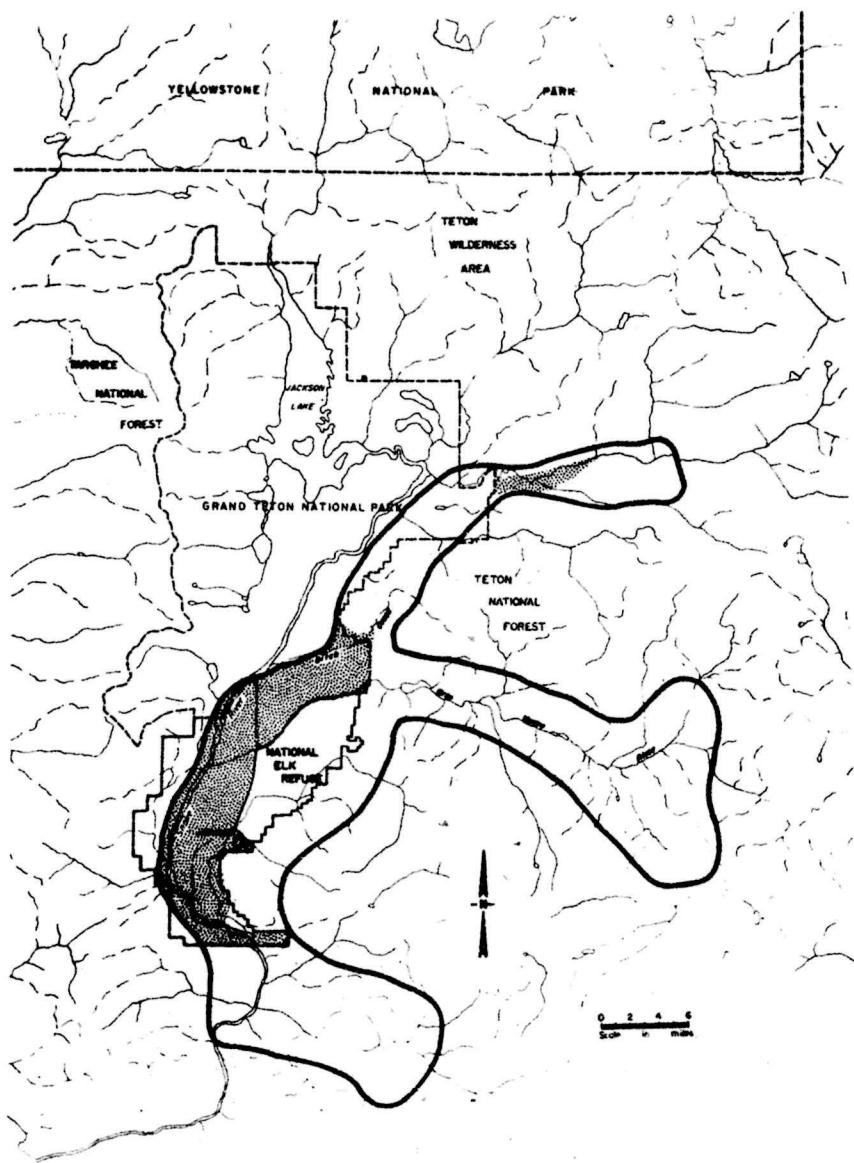


Fig. 6. Boundaries of historical elk winter range showing the main portion of Jackson Hole valley lands (shaded) where the animals are excluded because of human settlement or conflicts with agriculture.

of the herd now winters on or adjacent to feed grounds within the National Elk Refuge or other State feed grounds scattered through the south half of the Jackson Hole valley. The present Gros Ventre herd winters within the area mapped as their original range, but livestock grazing occurs on bottomlands and some slopes.

Relations to Green River Herd

Most literature on the Jackson Hole area relates a confusing story on the relationships between the Jackson Hole and the Green River elk herds. A review of this literature suggested how this may have occurred. Preble (1911) begins his report by stating that "Jackson Hole has long been the principal winter range for large numbers of elk." He later wrote, "In former years, large numbers, possibly the great majority of those that summer in the high ranges of northwestern Wyoming, wintered in the Red Desert." Graves and Nelson (1919) were probably influenced by this general statement when they reported, without supporting evidence, that "In early days elk passed through Jackson Valley to the plains beyond." Sheldon (1927) also did not present sources or evidence for his statement that: "Formerly herds of elk from the southern part of Yellowstone National Park . . . passed Jackson on their autumn migration and wintered in the Green River basin."

From this point on, a series of writers refer to some of this early literature as establishing that elk originally migrated through

the Jackson Hole area to the Green River basin (Allred, 1950; Murie, 1951; Craighead, 1952; Anderson, 1958). Supporting evidence for this migration is limited to observations of elk trails crossing divide areas between the Gros Ventre, Hoback, and Green River drainages and reports of large numbers of elk wintering in the Green River basin during the 1880's. Market hunting and starvation were reported to have greatly reduced the Green River winter herd by about 1913.

Barnes (1912) appears to be the first to specifically report that there were distinct Jackson Hole and Green River winter herds. This was based upon interviews with local Forest Service personnel and field observations. Brown (1947) reviewed some of the early literature and states that the "Green River migration (from Jackson Hole) is not established beyond dispute." Craighead (1952) and Anderson (1958) considered it unlikely that large numbers of Jackson Hole elk would suddenly cease to migrate because of settlement and hunting pressure in the distant Green River basin.

The confusion undoubtedly results from early unsupported statements being accepted as facts and some writers not being aware that summer segments from different winter herds may intermingle across drainage divides. It is this writer's judgment that summer herd segments from the Jackson Hole and Green River winter herds were intermingled on both sides of mountain divides between the Hoback, Gros Ventre, and Green

River drainages. Except for minor interchanges, the animals returned to their respective wintering areas. This migration and intermingling pattern occurs between animals from at least six different winter herds in southern Yellowstone Park today (see Intermingling between Herds.)

Past confusion concerning the relationships between the Jackson Hole and Green River herds has led to a general belief that Jackson Hole areas were not natural or historical winter ranges for any appreciable numbers of elk. It has also been suggested that the elk may have migrated out of Jackson Hole areas to a greater extent during severe winters. This would be difficult because mountain passes out of Jackson Hole would usually be blocked by the time snow depths finally force large numbers of elk into the valley. Migrations in anticipation of severe winters would be unlikely.

The 1887 to 1911 estimates of 15,000 to 25,000 elk in the Jackson Hole herd, with highest numbers reported during severe winters, should establish that the Jackson Hole and Gros Ventre valleys were historical winter areas. The first organized censuses, which accounted for approximately 20,000 animals, also seem to confirm that large numbers of elk were present historically.

Refuge Winter Herd

The average winter count of this herd was about 7,500 elk (4,233 to 9,804) during 13 of the winters from 1912 to 1954 (Anderson, 1958). An average of about 7,200 (5,746 to 11,017) was counted during

11 winters from 1955-56 to 1966-67 (Yorgason, writ. comm., 1968).

Figure 7 shows the estimated summer distribution of elk from the refuge winter herd during 1964. The 1963-64 winter herd of approximately 8,000 animals was assumed to increase by 23 percent and number about 9,800 during the summer.

Distribution estimates for south Yellowstone Park and adjoining Forest Service areas were primarily from counts of elk tracks crossing a 45-mile migration route transect. Migration trails in the snow were backtracked from an airplane to determine animal origins. Relative distributions within portions of the area were projected from repeated summer ground or aerial counts on sample areas.

Distribution estimates for Grand Teton mountain areas were obtained from track counts during fall migrations or as projections from ground, aerial, or photoelectric eye counts (on migration trails) within sample areas. Martinka (1965) obtained estimates for Grand Teton and refuge valley areas from ground counts and calculations that used sex and age or marked to unmarked animal ratios.

The estimates of 200 and 250 elk for Forest Service lands off the northeast corner of Grand Teton and the east side of the refuge were arbitrarily assigned. Repeated ground and aerial observations indicated that comparatively few elk summered in this area. The estimate of 400 elk for the Forest Service area east of Grand Teton was obtained from ground and aerial observations.

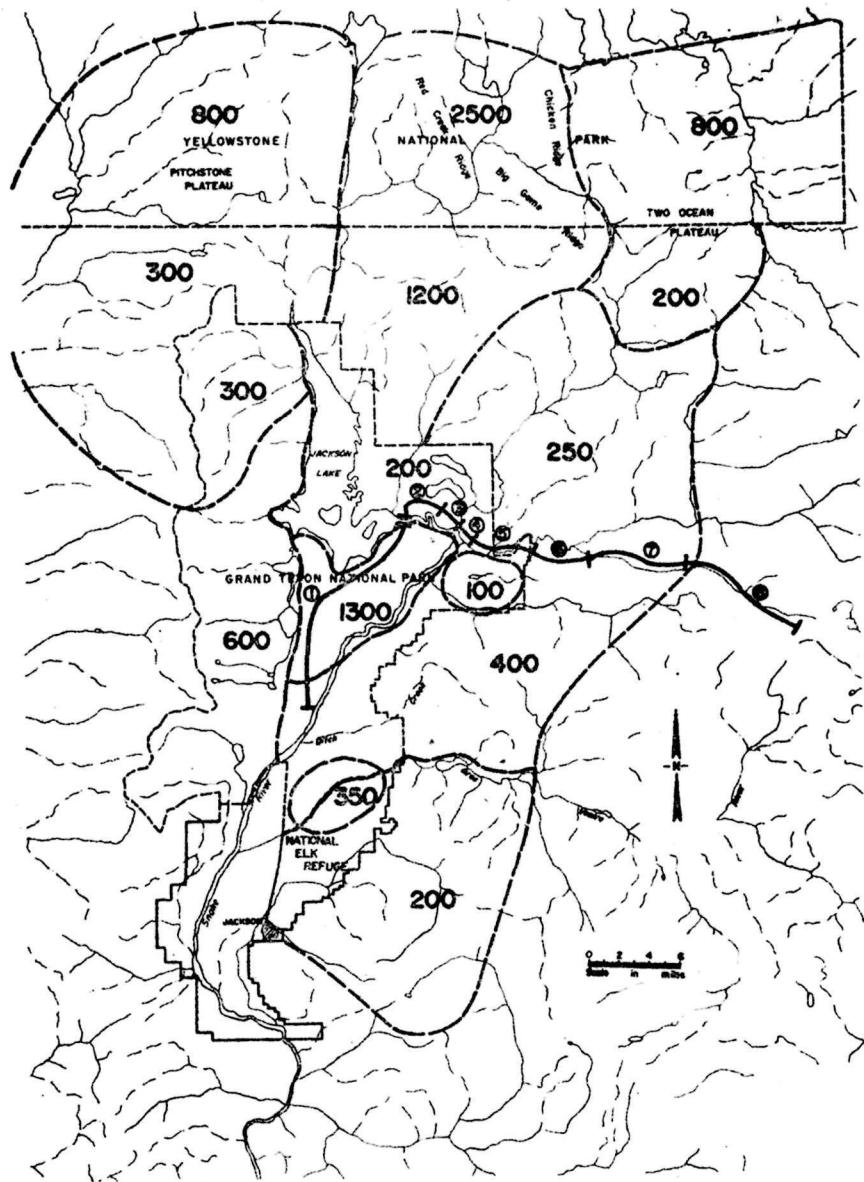


Fig. 7. Summer elk distributions by area and track count transects (—). Burnt Ridge, 1; Snake River, 2; Pacific Creek, 3; Buffalo River, 4; Park Boundary, 5; Blackrock, 6; Four Mile Meadow, 7; Togwotee Pass, 8.

The estimation methods applied over such an extensive area are admittedly rough. However, because unusually ideal fall track count conditions coincided with the greatest observed concentrations of elk on open subalpine summer ranges in 1964, they were the most complete measure of distributions obtained during the study.

National Park Elk

Figure 7 shows that about 70 percent of the elk originating from the refuge winter herd used summer ranges in the two national parks. As will be shown later, most of the animals migrating to and from the refuge also traveled over and used Grand Teton lands during spring and fall. Thus, the refuge winter herd was for the most part a national park herd during other seasons. This study therefore extended over the seasonal ranges and migratory routes used by the elk within the area shown on Figure 1.

Grand Teton

Figure 7 shows about 2,800 elk or 29 percent of the animals originating from the refuge winter herd summered on Grand Teton in 1964. About half of these summered on valley lands. Other elk from the refuge moved onto Grand Teton lands in late summer. Martinka's (1965) estimates of summer numbers in refuge and Grand Teton valley herds for 1963 and 1964 with subsequent estimates by the author and Houston (1968) are shown in Table 4. The comparatively low numbers in valley herds during 1965 and 1966 appeared to result from local shifts

Table 4.--Maximum late July or early August counts of elk on refuge and Grand Teton valley areas with calculations of probable numbers.

Year	Refuge	Grand Teton Valley	
	Maximum count	Maximum count	Probable number
1963	365	643	1,430 ¹
1964	346	632	1,400
1965	119	371	820 ¹
1966	101	401	890 ¹
1967	233	513	1,140 ¹

¹ Expanded using 1964 marked animal and sex ratio data (Martinka, 1965) that showed about 45 percent of the Grand Teton Valley Herd was observed by direct maximum counts.

of elk in response to high levels of disturbance associated with a forest insect control program (Cole, 1966 and 1967).

Osborne Russell's accounts from his 1834-1843 trips (Haines, 1965) suggest that appreciable numbers of elk originally summered in Jackson Hole valley areas. Between 1900 and about 1960, persons reporting on elk distributions either failed to mention valley summer herds in Grand Teton or stated that only small groups were present. McLaren (1966 pers. comm.), who has been a park ranger in Grand Teton since 1952, reported that he first noticed large groups of 100 or more elk summering in the park in 1960. An August 1962 aerial flight which only sampled valley areas accounted for 595 animals. A later October flight, which caught over 2,500 elk in the process of migrating from Grand Teton areas to the refuge, established that large numbers of elk were summering within the park and led to studies on valley and mountain herd segments.

Southern Yellowstone

Figure 7 shows about 4,100 elk, or 42 percent of the animals originating from the refuge winter herd, summered in southern Yellowstone. Preble (1911) estimated that approximately 4,000 summered in this region. The southern Yellowstone area shown to contain an estimated 2,500 elk in 1964 was the remote central mountain region where the greatest concentrations of summering elk were observed. It

was chosen for intensive study because it appeared to afford the best opportunity to assess the effects of large numbers of elk on high elevation summer ranges.

POPULATION DYNAMICS

This section is concerned with reproduction, sex and age structures, annual increase and mortality rates, and general population trends. The data relate primarily to animals that use or migrate over park lands, unless otherwise noted. Elk ages were determined by examining lower jaw teeth (Quimby and Gaab, 1957).

Female Reproduction

Breeding Ages

Examinations of reproductive organs from 751 elk killed on or adjacent to Grand Teton Park between October 1 and November 30 from 1962 through 1966 show most female elk become sexually mature at $2\frac{1}{2}$ years of age (Table 5). An average of about 15 percent of the yearling females were indicated to be sexually mature.

Pregnancy Rates

Table 5 also shows that about 78 percent of the females older than calves and 89 percent of the females older than yearlings were pregnant or had ovarian structures associated with fertility. These figures may approximate final pregnancy rates, despite some samples being from females killed before the end of the breeding season. The ages of 298 embryos (Morrison *et al.*, 1959) showed conception occurred from September 2 through November 29 from 1963 through 1966. About 16,

Table 5.--Numbers of different aged elk indicated to be pregnant or to have ovarian structures associated with fertility.

Ages	Sample	No. Pregnant ¹	No. Active ²	Pregnant or Active	
				No.	Pct.
½	23	0	0	0	0
1	108	4	12	16	15
2	106	73	21	94	89
3	123	72	40	112	91
4-7	259	155	85	230	89
8+	132	76	43	119	90
Totals	751	380	201	571	

¹ Embryo, embryonic tissue in uteri and/or corpora lutea 10 mm. or larger.

² Graffian follicles and/or partially lutenized follicles as well as corpora lutea with a diameter less than 10 mm. indicated ovarian activity associated with fertility.

62, and 22 percent of these conceptions occurred between September 2 through 20, September 21 through October 10, and October 11 through November 13 periods, respectively.

Murie (1951) examined 334 females older than yearlings killed on the refuge during December and January of 1935-36 and found 89 percent to be pregnant. Kittams (1953) and Greer (1966a) reported rates of 85 and 89 percent, respectively, for large samples of northern Yellowstone elk older than yearlings. The general agreement among figures suggests most estrous females become pregnant and animals older than yearlings have a fairly consistent high rate of pregnancy. The incidence of pregnant or "fertile" yearling females in this study varied, ranging from 0 to 27 percent in small yearly samples of 14 to 34 animals between 1963 and 1966. The above authors report yearling pregnancy rates that varied from 0 to 30 percent.

Relative Reproductive Success

Hunter reports for each of 907 female elk as "wet" or "dry" between 1963 and 1966 show about 47 percent of the females older than yearlings were lactating. About 10 percent of the 2-year-olds (bred as yearlings), 50 percent of the 3-year-olds, 58 percent of the 4-to-7-year-olds, and 49 percent of the 8-year-and-older animals in this sample were indicated to be nursing a calf. No yearlings were reported to be lactating.

Sex and Age Structures

Results from classifying elk on the refuge since the winter of 1927 are shown in Table 6. The average 1955-67 composition of 62 percent females, 19 percent calves, 6 percent yearling males and 13 percent adult males compares with averages of 64 percent females, 19 percent calves, 5 percent yearling males, and 12 percent adult males obtained over a 29-year period between 1927 and 1954. Some environmental regulation of population sex and age structures is suggested. Marked composition changes, such as occurred after the severe winter of 1961-62, were apparently caused by a proportionately greater late winter and spring mortality of adult male and calf elk and a subsequent reduction in the production of new calves. Records for this winter show that 34, 18, and 48 percent of 182 winter-killed animals examined were in the adult-yearling male, female, and calf classes, respectively.

Mortality and Increases

Newborn Mortality

Field observations indicated the peak of calving was between the last week of May and the first week of June. The difference between the 31 calves per 100 females in refuge winter herds since 1955 (Table 6) and the possible 78 percent pregnancy rate for females older than calves (Table 5) suggests that postnatal mortality averaged about 60 percent. The remaining 40 percent represented the realized calf production which averaged 19 percent of herd numbers by fall. Studies

Table 6.--Sex and age classifications of animals wintering on the National Elk Refuge.¹

Winter period	No. elk classified	Percent				Ratio/100 females	
		Females	Calves	Yearling	Adult	Calves	Males
1927-54 Average		64	19	5	12	30	26
1955-56	11,017	67	16	4	13	23	25
1958-59	5,595	56	22	4	18	39	40
1959-60	5,746	59	18	5	17	38	39
1960-61	6,701	59	20	7	14	34	35
1961-62	7,666	55	24	7	14	45	38
1962-63	5,827	70	13	6	11	18	24
1963-64	7,916	62	22	4	12	36	28
1964-65	7,946	62	19	7	12	31	31
1965-66	6,556	63	19	6	12	30	29
1966-67	7,369	63	21	5	11	33	17
1955-67 Average		62	19	6	13	31	31

¹ From National Elk Refuge records obtained by cooperative State and Federal agency classifications.

on elk by Greer (1966b), white-tailed deer (Odocoileus virginianus) by Verme (1962, 1963) and Murphy and Coates (1966) indicate that substantial mortality of young animals occurs at or shortly after birth. Young are born dead, too weak to nurse, or starve because of insufficient milk when females are in poor condition. Predation, accidents, and disease induced by adverse weather probably also remove proportionately greater numbers of young shortly after they are born.

Annual Increases and Other Mortality

The recent 10-year and earlier average of 19 percent calves in refuge winter herds show that yearly increases from winter to fall numbers could have averaged 23 percent (.19/.81 = .23). The more recent 4-year average of 20 percent calves since 1963-64 show a potential for 25 percent (.20/.80 = .25) increases. It was assumed that population numbers would remain relatively stable if hunting and other mortality removed numbers equal to calf production. Hunting mortality was considered to amount to the checked or calculated legal kill plus 10 percent (of the legal kill) for crippling losses and illegal removals. An 8 percent crippling, accidental death, and illegal kill loss was actually accounted for during the 1965 Grand Teton hunt (Cole, 1966).

Hunting mortality figures for the refuge winter herd could not be separated from those for the Gros Ventre herd before 1963-64. Combined figures were used to calculate yearly herd numbers which

gave the "best fit" with periodic censuses. A "best fit" was obtained over a 1955-63 period by using a 16 percent (.14/.86 = .16) increase rate (Table 7). The difference between expected and realized calf proportions (.19 - .14 = .05) suggested that late winter or spring mortality averaged about 5 percent of winter herds and approximately 4 percent of their subsequent fall numbers. Hunting mortality on fall herds removed an additional 16 percent. This total 20 percent mortality apparently removed numbers greater than those being added by calf production. Numbers in the two herds appear to have declined from about 14,700 to 12,600 animals over the period. The difference amounted to about 260 animals per year or roughly 2 percent of winter numbers. Similar declines occurred over the next 4 years with hunting removals that averaged about 21 percent.

Hunting mortality on the refuge winter herd probably averaged about 19 percent of fall herd numbers since 1963-64 (Table 8). The differences between censuses and calculated 1966-67 herd sizes and the 20 percent potential could represent sampling errors or indications that additional mortality from causes other than hunting continued. The latter was probably the case. Differences between potential and observed ratios of yearling male elk in July classifications ($N = 3,717$) from mountain areas indicated that overwinter and spring mortality on calf elk was about 19 percent or equal to 3 percent of winter herd numbers after the deep snow winter of 1964-65; not

Table 7.--Population size and hunter kill information from combined refuge and Gros Ventre herds.¹

Years	Winter herd size ²	Calculated fall herd	Hunter kill	
			Number	Percent
1955-63	12,800 Avg.	15,200 Avg.	2,400 Avg.	16
1963-64	11,500	14,600	3,100	21
1964-65	11,300	14,400	3,100	21
1965-66	10,500	13,800	3,300	24
1966-67	10,600	13,100	2,500	19

¹ The 1955-56 population figure of 14,685 and hunter kill records from Wyoming Game and Fish Commission reports.

² Calculated using 16 percent annual increase up to 1963-64 and 25 percent thereafter. The 1966-67 calculated population was approximately 100 animals above a subsequent census.

Table 8.--Population size and hunter kill information from refuge winter herd.

Years	Winter herd size ¹	Calculated fall herd	Hunter kills ²	
			Number	Percent
1963-64	7,900 (7,916)	9,400	1,500	16
1964-65	8,000 (8,096)	9,900	1,900	19
1965-66	7,600 (6,556)	10,000	2,400	24
1966-67	7,900 (7,663)	9,500	1,600	17

¹ Calculated using 25 percent increase rates with censused numbers in ().

² Proportion calculations using elk checked at park and state check stations.

detectable following the mild or average 1962-63, 1963-64 and 1965-66 winters. Recorded mortality of all sex and age classes on refuge feed grounds was 1 percent or less of winter numbers during these years (Yorgason, 1968 writ. comm.). Comparable calculations indicated calf mortality during and after the severe 1961-62 winter was about 75 percent or equal to 18 percent of winter herd numbers. Only a 2 percent mortality of all sex and ages was accounted for on feed grounds. Ratio calculations used calf per 100 female values in winter herds as recruitment potentials and adjusted observed yearling per 100 female values (Y) as $2Y/100-Y$. Equal calf sex ratios with proportionate mortality and either proportionate or relatively low adult female mortality was assumed.

The calf mortality figures could be slightly inflated because of a tendency for proportionately greater numbers of yearlings to remain in valley areas (Martinka, 1965). However, it seems probable that a 2 to 3 percent mortality of both calf and adult elk could be expected during most winters and at least 15 percent for severe winters. Mortality that averaged at least 5 percent over mild, average, and severe winters may have been partially density-independent with herd sizes in the 6,000 to 8,000 range.

Population Trends

General trends for the refuge winter herd and the Jackson Hole herd as a whole are shown with available hunter harvest data on

Figure 8. Population trend figures were from highest censuses (N = 11) obtained within 5-year periods since 1911. These were usually obtained during more severe winters. Census and/or harvest records were from Preble (1911), Sheldon (1927), Anderson (1958), and Yorgason (1968 writ. comm.).

Population trends for the refuge herd appear to have been relatively "stable" for the past 57 years. Winter herd numbers have apparently fluctuated within a 6,000 to 8,000 range about 65 percent of this time, a 5,000 to 9,000 range 78 percent of the time, and a 5,000 to 10,000 range with a 98 percent frequency (Table 9). Trends and fluctuations showed no obvious relationship to hunting removals until after 1950 (Figure 8).

After 1950, higher hunting removals (from a smaller total population) may have partially reduced the extreme low fluctuations in herd numbers. Periodic upward fluctuations continued through the 1950's. Acquisitions of additional portions of the elk winter range by and after 1950 could have also contributed to these, as well as to the reduced downward fluctuations.

Figure 8 shows population trends for the Jackson Hole herd as a whole have been downward since 1911. The general declines and fluctuations up to about 1941 were undoubtedly caused by some thing(s) other than the relatively low hunting removals. Reports by Preble

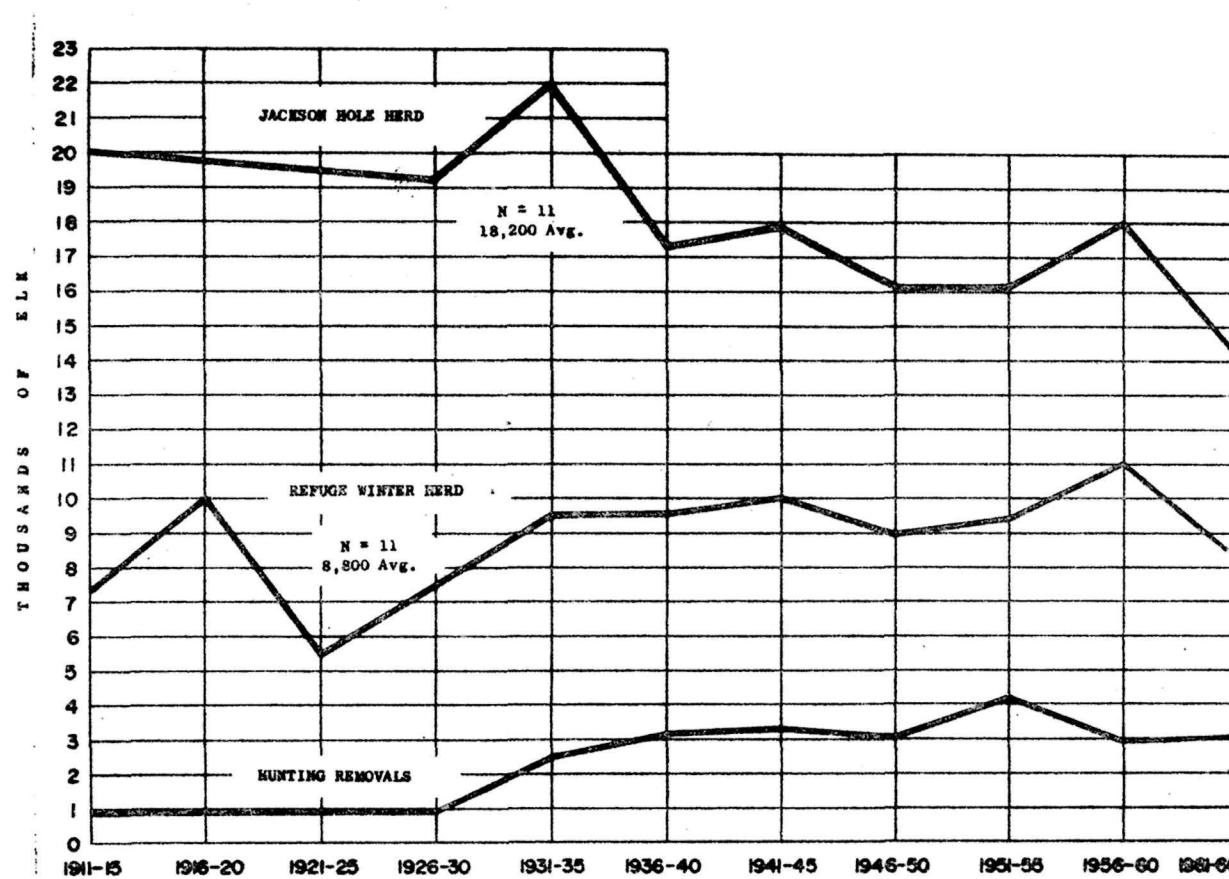


Fig. 8. Population trends of the Jackson Hole herd and refuge winter segment, as indicated by maximum counts and/or estimates within periods since 1911 and corresponding average hunting removals.

Table 9.--Frequency with which refuge winter herd occurred within different size ranges during 54 years since 1912.

Winter herd size	Tabulated frequency		Probable frequency ¹	
	No. years	Percent	No. years	Percent
6,000 to 8,000	30	56	35	65
5,000 to 9,000	36	61	42	78
5,000 to 10,000	44	81	53	98
5,000 to 11,000	45	83	54	100

¹ Interspersed counts of 3,000 to 4,500 animals during 9 years with milder winters and/or population lows proportionately assigned within the different size ranges.

(1911), Barnes (1912), Graves and Nelson (1919), Sheldon (1927), and Anderson (1958) suggest that livestock grazing and agricultural developments may have progressively reduced winter food for elk within this period.

Enactment of a State game damage law in 1939 undoubtedly gave impetus to securing increased hunting removals from herd segments conflicting with agriculture. These animals, elk in scattered wintering areas off established feed grounds and groups that migrated through areas where they were particularly vulnerable, appear to have been the herd segments mainly reduced by the sustained high average hunting removals during the late 1940's and the 1950's (see Fall Migrations). Censuses of elk in scattered groups off established feed grounds ceased in 1959, but it is probably safe to assume that between 13,000 and 15,000 elk have occurred in the Jackson Hole herd through the 1960's.

While hunting removals could be considered the cause for the general population declines in the Jackson Hole elk herd, the actual agents were human settlement and agricultural developments on the animals' historical winter range. The refuge herd, which had essential portions of its original winter range restored, appears to have maintained its numbers while other segments declined. Winter herds in the southern portions of the Jackson Hole valley were artificially fed on small areas to hold the animals off extensive bottomlands that

have become hayfields and livestock pastures. Such feeding may have substituted for the loss of bottomland foraging areas, but not without consequences on herd numbers (see Ecology).

The overall decline in the Jackson Hole elk herd probably approximates 30 percent of their 1910-11 numbers. This coincides with the animals being restricted from using about one-third of their original winter ranges within the main Jackson Hole valley and substituting domestic stock for elk grazing on other wintering areas. Other human influences additionally restricted the animals' winter distributions and use of food sources with adverse population consequences. As will be discussed in later sections, the availability of a complex of interspersed bottomland, swale, and slope sites may be as important in maintaining elk numbers as some total acreage of winter range.

ELK HABITS

Information on elk habits was obtained while covering established routes through the study area. Routes were covered under a system which employed periodic aerial flights, horseback and foot travel through roadless areas, and vehicle trips. Observations of elk numbers, activities, occurrences on different habitat types, and locations were recorded. Most observations were made during early morning or evening when the animals were more observable and feeding. Animal locations were determined from gridded topographic maps. Sex and age classifications and identifications of individually marked animals were made whenever possible. Binoculars and a variable 15 to 60 power spotting scope were used.

Movements

Spring Migrations

During January through March periods when they were being fed hay, large groups of 1,000 to 4,000 elk concentrated on two or three feed ground sites within the south half of the National Elk Refuge. As snow-free areas became available and new grass growth started during April and May, the animals dispersed from feed grounds and ranged into the north half of the refuge. Snowstorms or inclement weather halted or even reversed these spring dispersals.

First migrations of significant numbers of elk (500 or more counted) off the refuge and onto Grand Teton lands occurred on April 30 in 1963, May 14 in 1964 and 1965, May 2 in 1966, and May 9 in 1967. Numbers moving onto Grand Teton Park progressively increased once migrations started. Movements to limited snow-free areas in the northern portion of the valley occurred over extensive flats with up to 1 foot of snow and over much deeper drifts. Such movements took the animals from refuge areas where new vegetation growth was abundant into areas where new growth either had not started or was so short as to be unavailable.

The rate of spring movements by elk which migrated through Grand Teton into southern Yellowstone and the use of particular areas by females for calving appeared to be strongly influenced by snow accumulations in mountain passes. Proportionately fewer females with calves appeared to be present in southern Yellowstone during late June in 1962 and 1965 when June 1 snow depths at a representative 8,000 foot mountain pass exceeded 60 inches (Table 10). The proportion doubled in 1963 and 1966 when June 1 snow depths were 28 and 20 inches respectively. The highest proportion in 1964, with a 43-inch snow depth, appeared to represent a situation where intermediate mountain areas had sufficient snow accumulations to hasten, but not prevent, movements to lower elevations in Yellowstone Park. Yorgason (1964) reported from aerial flight observations that large numbers of elk migrated into Yellowstone Park over at least 10 continuous miles of snow cover

Table 10.--Numbers and sex and age classes of elk observed in Yellowstone Park areas in late June in relation to snow depths at an 8,000 foot station.

Period	Total elk observed	Classi- fied	Percentages			Calves/100 other adults	June 1 snow depths
			Adult males	Other adults	Calves		
June 24, 28, 1962	268	241	30	63	7	10	61
June 24-27, 1963	571	449	25	62	13	20	28
June 26, 1964	673	476	14	65	21	32	43
June 29, 1965	408	380	34	60	6	10	70
June 27-July 1, 1966	872	781	15	71	14	20	20

between May 9 and June 13. Alternative, more snow-free, but less direct, routes up main drainage courses were available. These apparently did not serve as alternative migratory routes for the majority of the animals that habitually traveled through mountain areas.

Elevational Movements

Recorded locations from 16,103 elk observations showed that progressively greater numbers of animals moved from low to high elevation ranges above 8,500 feet in southern Yellowstone through June and July (Table 11). Peak numbers were observed on high ranges during the last half of July during 4 of the 5 years; during the first half of July in 1966. The animals used forest types at both high and low elevations to a greater extent during the first half of August and were generally less observable. Movements from high to low elevations and even greater use of forest cover occurred during the last half of August and September. Some animals moved back to high elevations during late September and October in years when fall storms were not too severe.

Sex and Age Differences

Classifications of 12,682 elk showed adult males and females not attending calves were most involved in initial June movements onto high elevation ranges above 8,500 feet (Figure 9). Females with calves or attached to female-calf groups was the population segment most involved in progressive movements to high ranges through July. Movements of adult males to high ranges appeared to be largely completed

Table 11.--Numbers and percentages of elk observed on mountain range areas above 8,500 feet, 1962-1966.

Period	Total elk observed	Elk above 8,500 feet	
		Number	Percent
June 16-30	3,372	1,525	45
July 1-15	3,126	1,824	58
July 16-31	4,606	3,597	78
August 1-15	2,521	1,800	71
August 16-31	1,217	400	33
September and October	<u>1,261</u>	<u>485</u>	38
	16,103	9,631	

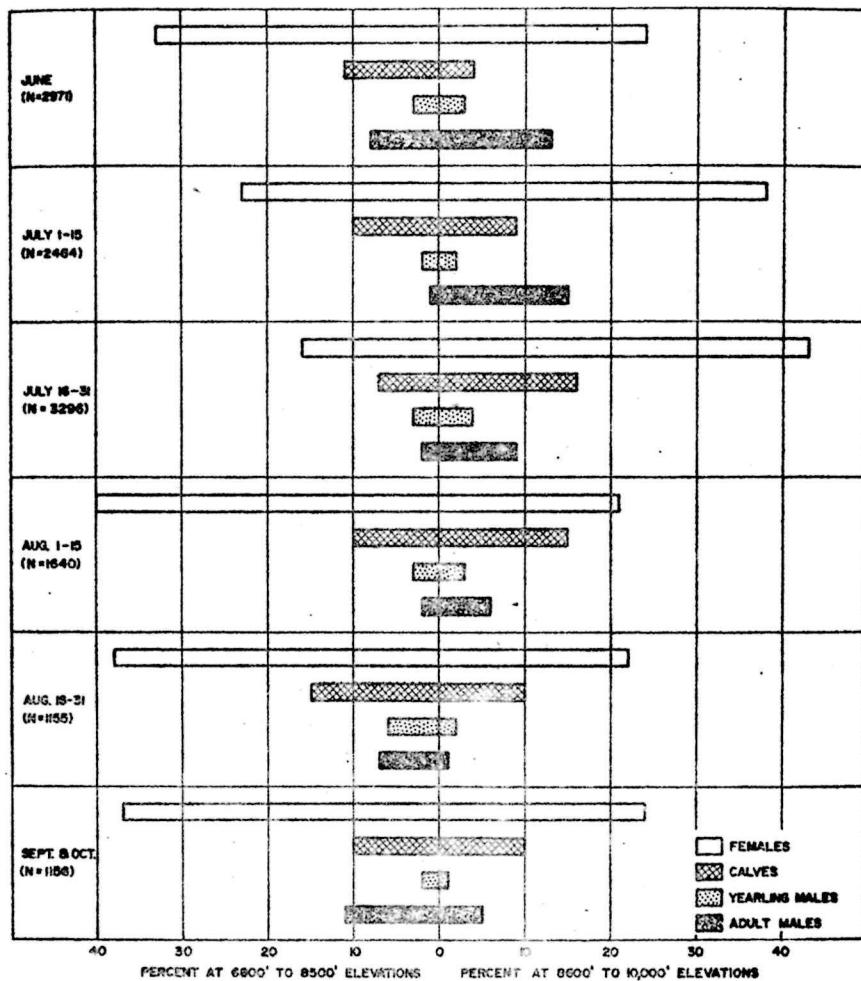


Fig. 9. Percentages of female, calf, yearling, and adult male elk observed between 6800 to 9800 and 8600 to 10,000 ft. elevations within mountain areas during June through October periods, 1962-1966.

by mid-July. Yearling males occurred in about equal proportions in early and late arriving groups. Adult males usually occurred in small groups apart from other population segments. and at the highest elevations during late July. All sex and age classes were involved in early or mid-August dispersals into forest habitats and late August through September movements to lower elevations.

Insect Relationships

Midsummer aggregations of large numbers of elk at high elevations were influenced by molesting insects. Horseflies (Tabanidae) were obvious molesters of elk. Mosquitoes (Culicidae) caused avoidance reactions when very abundant. An unknown insect, possibly a heel fly (Oestridae), caused pronounced avoidance reactions in elk, but did not molest men or horses.

Numbers of elk observed on high elevation ranges during July and early August with different intensities of insect molesting activity are shown in Table 12. Intensities were considered high when elk were almost solely occupied with avoiding or dislodging insects, moderate when such actions were secondary to other activities, and light when the animals appeared to be completely or almost unmolested. Peaks of molesting activity were through midday periods. High intensities caused obvious aggregations of elk on ridgetops and/or bedding in dense vegetation.

Table 12.--Numbers and average size of elk groups observed on high elevation ranges in relation to molesting insects.

Year	Relative abundance of molesting insects	Elk above 8,500 feet		
		Number and average group size () July 1-15	July 16-31	August 1-5
1962	Moderate late July, early August	328 (27)	650 (20)	609 (24)
1963	Intense late July; moderate early August	422 (23)	1,036 (77)	486 (20)
1964	Low all summer	256 (21)	671 (15)	136 (8)
1965	Intense early August	50 (27)	990 (35)	493 (19)
1966	Intense early July	768 (18)	250 (10)	76 (7)

Effects of molesting insects were most apparent in late July of 1963 and early July of 1966. This coincided with relatively high numbers of tabanids and the probable heel fly. The intense molesting action in early July caused the unusual occurrence of some females with young calves at elevations up to 10,000 feet. The habits of elk with the almost complete absence of insect molesting activity are reflected by the 1964 data. Elk remained scattered and in comparatively small groups throughout the summer. Similar scattered distributions occurred after early July of 1966.

Dispersals of elk into smaller groups usually occurred in early August. The effects of insect molesting activity occurring in or extending into August are illustrated by the 1962 and 1965 data. This suggests insects partially prevented elk dispersals into smaller groups. The relatively high elk numbers and group sizes seen during late July of 1965 appeared to result from the first migrations of large numbers of females with calves into Yellowstone National Park. These migratory groups dispersed and remained scattered in the absence of significant numbers of molesting insects through the remainder of the summer.

Intermingling Between Herds

The extent of intermingling between animals from the refuge and northern Yellowstone winter herds on the study area was indicated by 2,340 locations of marked elk (Table 13). Animals from these two

Table 13.--Numbers of marked refuge and Northern Yellowstone elk observed within distance zones from the south boundary of the National Elk Refuge, 1963-1966.

Month	Grand Teton	North of	Southern Yellowstone Park	
	valley 7-25 miles	valley 26-44 miles	45-49 miles	50-55 miles
April	30			
May	1,369 1 NY	33		
June	198	44	23	17 4 NY
July		39	93 3 NY	133 19 NY
August			45 1 NY	50 6 NY
September		11	19	7 1 NY
October	97 1 NY	7	1	
November	45	43		
Refuge total	1,739	177	181	207
Northern Yellowstone total	2		4	30

almost 100-mile distant winter areas were shown to be intermingled to the greatest extent on Yellowstone summer ranges 50 to 55 miles from the south boundary of the refuge, to a lesser extent within 45 to 49 miles, and to a very limited extent on ranges south of Yellowstone Park. Simple proportion calculations, that related estimated herd sizes to numbers marked, suggested that at least 90 percent of the animals using the sampled central southern Yellowstone ranges were from the refuge winter herd during 1963 and 1964 (Cole, 1965). This figure may not be entirely applicable for the east and west portions of southern Yellowstone Park. A portion of the animals summering in the eastern Two Ocean Plateau and Thorofare regions were from the Gros Ventre winter herd and another herd which migrated east to winter ranges near Dubois, Wyoming (Murie, 1929; Anderson, 1958; Yorgason, 1964). A portion of the animals summering on the western Pitchstone Plateau region migrated southwest to Idaho winter ranges (Anderson, 1958) and possibly north to winter ranges along the Firehole and Madison Rivers inside Yellowstone Park.

In summary, the intermingling of elk from widely separated winter ranges, on and across mountain divide or plateau areas in southern Yellowstone Park, was a common occurrence. The presence of marked animals from two different winter ranges in a single group was also common during this study. Marked animals from three different winter ranges were occasionally seen together.

Despite this common intermingling on summer ranges, interchanges of animals between winter herds seemed to be less than would be expected. Over the 1962 through 1967 period, a maximum of three marked animals from the northern Yellowstone herd were present on refuge winter ranges during the winter of 1964-65. Numbers observed during other years ranged from none to two. Two marked refuge elk were observed on northern Yellowstone winter ranges in 1963-64; single animals, during each of the other 4 years.

Fall Migrations

The general pattern of fall and early winter migrations back to the National Elk Refuge is illustrated by the September through November locations of marked elk (Table 13). Other migration information was primarily from track counts on a 47-mile road transect and periodic counts of animals appearing on the refuge. The road transect started at park headquarters at Moose, extended through the west side of the park to Moran, and from here east to the top of Togwotee Pass (Figure 7). Elk track counts were started in 1945 by Cahalane (1949). They have been made cooperatively with Wyoming personnel since 1950. Park Service and Wyoming biologists alternate senior authorship in preparing annual migration reports. Interested persons are referred to these reports for detailed information on migrations.

Migration Chronology

Records over a 40-year (1927-1966) period date main migrations as occurring 4 times in October, 24 times in November, and 2 times in December (Yorgason and Cole, 1967). Three of the four "main" October migrations occurred consecutively from 1960 to 1962. Combined numbers migrating during November and December were actually greater than October. Track count records and aerial observations indicated most of the elk making early October migrations came from Grand Teton valley and mountain areas. Refuge counts indicate habitual October migrations by comparatively small numbers of elk started in 1958 (Table 14). Road closures within Grand Teton Park and specially designed hunting seasons aided in curtailing early migrations after 1964.

Tabulations of tracks crossing Grand Teton transects after snow-fall and direct counts of elk within periods show the overall chronology of migrations for the majority of the animals summering in both national parks. Nine counts over the 10 years between 1957 and 1966 indicated that, on the average, about 5, 16, 28, 40, and 11 percent of these elk had migrated during October 1-15, October 16-31, November 1-15, November 16-30, and December 1-30 periods, respectively. These yearly track counts sampled the movements of an estimated 5,000 to 7,000 animals, depending upon the extent of early migrations before snow cover. Counts tallied between 40 and 80 percent of these animals. Low counts mainly resulted from snowstorms or mass movements obscuring tracks, or thaws that temporarily removed snow cover.

Table 14.--Bimonthly counts of elk on the south two-thirds of the National Elk Refuge.

	1958	1959	1960	1961	1962	1963	1964	1965	1966
Oct. 1-15	300	775	1,100	1,726	2,450	145	145	0	36
Oct. 16-31	809	1,550	2,500	3,470	2,500	728	1,000	375	300
Nov. 1-15	1,500	2,533	3,500	4,500	2,490	5,000	1,800	521	1,540
Nov. 16-30	5,000	4,000	5,800	-- ¹	2,600	7,000	2,900	2,300	4,000
Dec. 1-15	4,200	3,800	5,850	6,200	-- ¹	6,225	4,500	3,900	-- ¹
Dec. 16-31	4,500	4,200	6,200	-- ¹	4,036	-- ¹	6,000	-- ¹	-- ¹

¹ Elk scattered; count not possible.

Number Relationships

Accumulated track count records since 1949 show that the numbers and proportions of elk crossing transect areas inside and outside Grand Teton have greatly changed over an 18-year period. Illustrative counts, at selected 6- and 7-year intervals, which sampled large numbers of elk are shown in Table 15. With the exception of the most eastern portion of the Four Mile Meadow to Togwotee Pass section (Figure 7), the transect mainly sampled animals migrating to the National Elk Refuge. The changes in numbers and proportions crossing different transect sections were generally progressive from about 1950 through 1958 and coincided with a period of relatively high hunting kills. Proportionately greater kills were apparently made on the more accessible migratory segments that crossed the eastern portions of Grand Teton (Buffalo and park boundary sections) and the transect sections outside park boundaries. Numbers decreased. Migratory segments that traveled through less accessible hunting areas (roadless or roads blocked by snow when migrations occurred) north of Grand Teton before crossing the Snake River and Pacific Creek transect sections were obviously less heavily hunted. These migratory segments increased. Increases in elk numbers crossing the Burnt Ridge section resulted from a buildup of a resident summer herd in Grand Teton after hunting ceased on low security level valley habitats west of the Snake River.

In summary, the increases and decreases in particular migratory segments appeared to be in relation to their accessibility to hunters.

Table 15.--Numbers of elk tracks crossing transects inside and outside Grand Teton National Park in years of maximum counts at 6 to 7 year intervals.

	1949	1957	1964
Park transects:			
Burnt Ridge	424	957	1,672
Snake River	1,285	1,091	2,122
Pacific Creek Bridge	700	1,025	1,416
Buffalo River Bridge	266	259	99
Park Boundary	<u>842</u>	<u>626</u>	<u>371</u>
Park total	3,517	3,958	5,680
Outside transects:			
Blackrock	3,492	1,159	395
Four Mile Meadow	3,146	1,112	1,028
Togwotee Pass	<u>700</u>	<u>3,102</u>	<u>470</u>
Outside total	7,338	5,373	1,893
Grand total	10,855	9,331	7,573
Percent in park	32	42	75
Percent outside park	68	58	25

Accessibility was largely determined by the presence of roads or the extent to which roads were blocked by snow during elk migrations.

Habitat Use

Patterns of elk habitat use within Grand Teton and refuge valley areas and mountain areas extending into southern Yellowstone Park were determined by recording the numbers of animals observed on the different vegetation types while covering routes. Multiple records were made when undisturbed animals moved from one type to another. Animals were about equally observable when they were using all but the forest types. Numbers observed in the forest types during the August through October period of greatest use were weighted by assigning the difference from maximum July counts to these types.

Valley Areas

General patterns of habitat use within valley areas were shown by 82,223 recorded elk observations between April and December (Figure 10). The sample included 42,237 May through October observations from 1963 and 1964 that were made by Martinka (1965). Additional observations were obtained within and on either side of this period from 1963 through 1966. The average sample size for monthly periods was 9,136 (5,525 to 19,398). April, November and December averages were from selected samples when large numbers of elk were freeranging off refuge feed grounds.

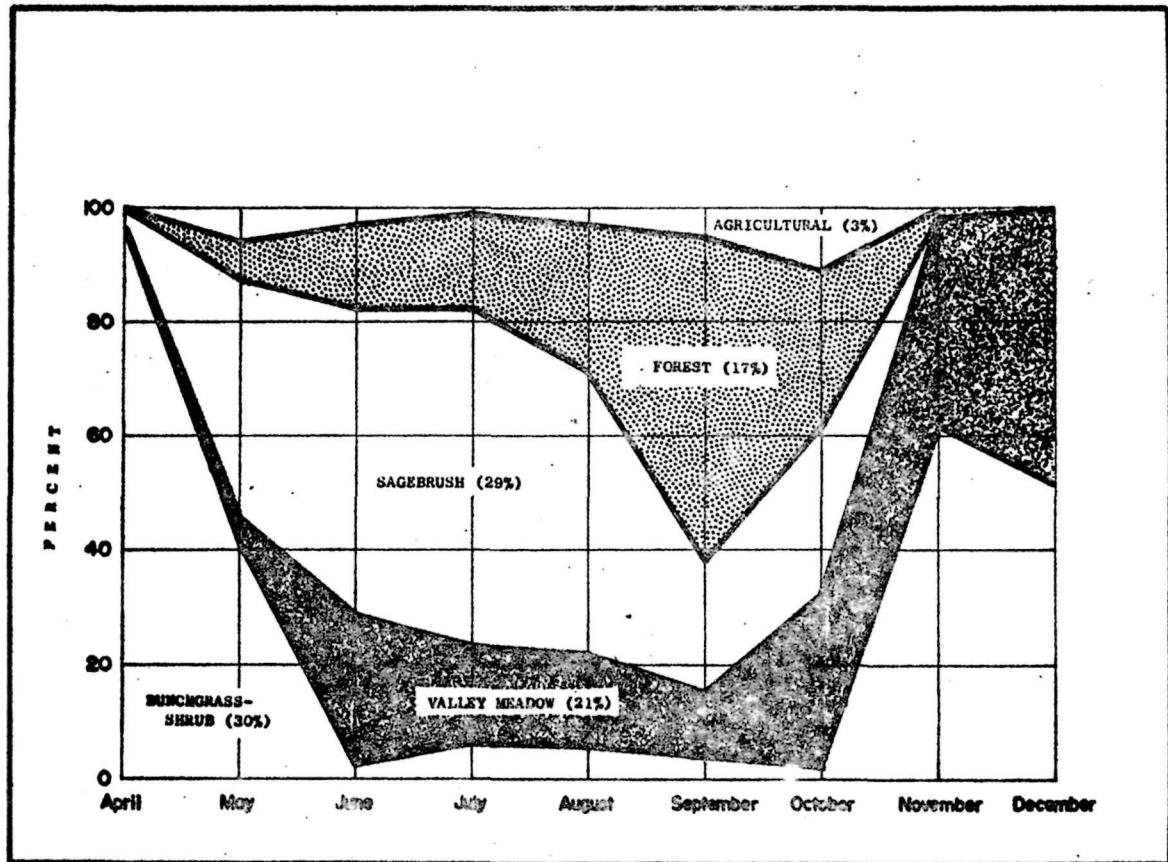


Fig. 10. Relative use of different vegetation types in valley areas as indicated from observations of 82,223 elk between April and December, 1962-1966.

The recorded observations from late April illustrate the animals' maximum use of the bunchgrass-shrub type on upland slope sites after they dispersed from feed grounds and ranged over the northern portions of the refuge. Extensive areas of the type on an alluvial fan and hayfields were used to a greater extent when the animals first moved off feed grounds. May observations illustrate the animals' progressive shift to the sagebrush type as they migrated onto and used valley floor areas in Grand Teton Park. The bunchgrass-shrub type on south slopes remained important for its more advanced stages of vegetation growth. The predominant use of the sagebrush type for foraging is shown by the June through August observations.

Shifts to greater use of forest types occurred in August and increased during September. The October observations showed the animals increased their use of the valley meadow type and hayfields. This mainly occurred in October of 1962 and 1963 when large numbers of elk migrated to the refuge. Restrictions of human disturbances in Grand Teton areas closed to hunting and special hunting seasons reduced these movements in subsequent years and led to the greater use of the sagebrush type within the park.

November and December observations illustrate the animals' use of extensive bunchgrass-shrub and valley meadow types. Use at this time of year appeared to be almost exclusively on high producing

bottomland sites or within swales on uplands. Variable additional use occurred on hayfields in some years. This appeared to depend upon whether fall rains caused regrowth or they had been left uncut. The start of artificial feeding in January caused the majority of the animals on the refuge to abruptly leave natural food sources and move onto feed grounds.

Mountain Areas

Patterns of use on different vegetation types in mountain areas were shown by 20,017 recorded elk observations between June and October of 1962 through 1966 (Figure 11). Sample sizes for periods ranged from 641 during the first half of June to 6,543 during the last half of July.

Figure 11 shows that the majority of the animals arriving in mountain areas in early June used the relatively low elevation valley meadow (willow-sedge stage) and forest park types. These were principally female-calf groups. Late June through July observations illustrate progressive shifts in elk use from the low elevation types to the intermediate elevation burn type and finally, the herbland type. Maximum use of the highest elevation subalpine meadow type usually occurred during late July. August through October observations illustrate the elks' dispersals into and predominant use of forest types.

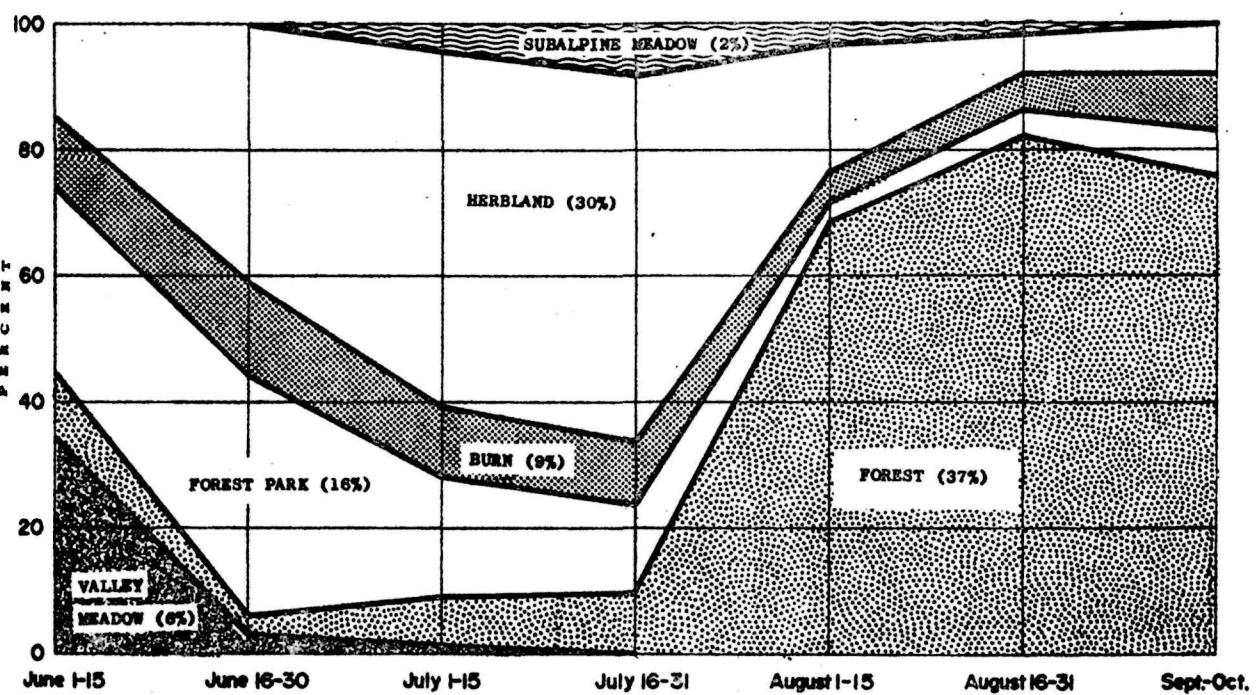


Fig. 11. Relative use of different vegetation types in mountain areas as indicated from observations of 20,017 elk between June and October, 1962-66.

In 1962, a late emergence of molesting insects appeared to hold substantial numbers of elk on the hermland and subalpine meadow types into the first half of August (Table 12). In late July and early August of 1964, molesting insects were scarce. Elk occurred in scattered small groups and used forest and low elevation forest park types to the greatest extent observed during the study. In 1965, snow accumulations along migratory routes appeared to preclude the usual use of lower elevation mountain areas for calving. Large female-calf groups moved directly onto Yellowstone's high elevation hermland types in late July. In 1966, an early July emergence of molesting insects caused concentrations on high elevation herblands comparable to late July and early August highs observed in other years. Subsequent late July and early August use of high elevation ranges was the lowest observed during the study.

Food Habits

Information on food habits was obtained by recording 262,602 instances of plant use at 473 elk feeding sites. The unit for recording one instance of use from April 16 through November 15 was a rooted stem for forbs and single-stemmed grasses, a leader for shrubs, or a distinct clump of stems for bunchgrasses. A unit recorded during this period could be considered roughly equivalent to a "bite."

During the November 16 through April 15 winter period, elk frequently grazed bunchgrasses closely. On sites where shrubs were also used the above system was modified. Recorded instances of use on bluebunch wheatgrass, or other grasses with a similar life form that were grazed to average heights of 1, 2, 3, or 4 inches, were multiplied by weight values of 25, 20, 15 or 10, respectively. Instances of use on needle-and-thread, or other comparable grasses that were grazed to 0.5, 1, or 2 inch average heights, were multiplied by weight values of 11, 8, or 5, respectively. These weight values made units of use on bunchgrass plants more equivalent in weight to units for shrubs and the few forbs used during this period.

The base weight values of 25 and 11 were obtained by dividing the average air dry weight of 100 2-inch shrub leaders ($N = 400$) into the air dry weight of 100 bluebunch wheatgrass plants clipped to 1 inch and 100 needle-and-thread plants clipped to 0.5 inch. Leader samples were from Douglas rabbitbrush and winterfat and weights were approximately equal. Weight values for other bunchgrass stubble heights were obtained from proportion calculations that utilized the "utilization gauge" developed by Lomason and Jenson (1938).

The only departure from the described method was that the side branches of tall larkspur and individual stems of giant wild rye plants were considered special units for recording one unit of use. Most

feeding sites were examined while fresh use could be detected, but some were allowed to accumulate use over known periods when elk were the only animals involved. Examinations were purposely restricted to sites that had not been used to the extent that animal preference was obscured.

Elk use was mainly on cured grass during the November 16 through April 15 winter season. New growth of grasses usually became increasingly available as a food source through the first half of spring; forbs, during the last half of the period (April 16 to May 16 or June 30). Variation occurred with late and early springs. Forbs were readily available on most habitat types through the summer (up to September 15), though some cured to the extent that they ceased to be used. The freezing or curing of forbs in fall (September 16 to November 15) resulted in elk shifting their use to other forage sources.

Plant terminology generally follows Davis (1952) and Booth and Wright (1959), except that grasslike sedge and rush species are included in the grass forage class. Grazed and vegetative forms of fleabane or aster plants could not be readily identified and are collectively called asters. Instances of elk use on plant species (or related groups) were calculated as percentages of the total recorded use at each feeding site. Percentages were averaged for the different vegetation types by seasonal periods. Results by forage classes are shown in

Table 16. Plants that received 5 percent or more of the recorded use and/or were used at 25 percent or more of the sites are shown in Appendix Tables V through IX. An alphabetical listing of common and scientific plant names is given in Appendix I. Discussions of food habits by seasons follow.

Winter

Grasses were the predominant class of forage used by elk feeding on the upland bunchgrass-shrub and valley meadow types (Table 16). Bluebunch wheatgrass and Douglas rabbitbrush were the two most important plant species used on the upland type as a whole (Appendix V). Needle-and-thread or bluegrasses were additionally important food items within swales during the first half of winter and again on a variety of other upland sites as new growth started during late winter. Combined use on these four plant species averaged 79 percent of the recorded instances of use. Sedge, bluegrass, and willow were the most important species used on the valley meadow type (Appendix VI). Combined use averaged 76 percent.

Shrubs were the predominant class of forage used by elk feeding within forest types during winter. Willow, narrowleaf cottonwood, aspen, and silverberry in combination with bluegrasses were the most important items used in aspen or cottonwood stands (Appendix VII). Combined use was 70 percent. A variety of shrubs and sedge served

Table 16.--Instances of plant use by forage class at 473 elk feeding sites on different vegetation types during seasonal periods.¹

Seasons ²	Vegetation types	No. sites	Instances of use	Percent		
				Grass ³	Forbs	Shrubs
Winter	Bunchgrass-Shrub	51	163,837	72	2	26
	Valley-Meadow	8	6,152	71	2	27
	Forest (Deciduous)	4	3,016	43	6	51
	Forest (Coniferous)	9	9,657	37	9	54
Spring	Bunchgrass-Shrub	24	4,606	75	14	11
	Valley-Meadow	19	7,836	60	36	4
	Sagebrush	43	8,257	54	44	2
	Forest Park	37	8,107	33	66	1
	Burn	5	1,422	57	42	1
	Herbland	5	1,102	48	52	0
	Forest (Deciduous)	16	4,127	50	49	1
	Forest (Coniferous)	18	3,053	40	39	21
Summer	Valley-Meadow	9	1,484	23	7	70
	Sagebrush	59	7,637	4	92	4
	Forest Park	33	5,369	14	82	4
	Burn	15	2,546	13	83	4
	Herbland	30	5,488	15	85	0
	Subalpine Meadow	14	3,351	50	50	0
	Forest (Deciduous)	17	3,256	25	50	25
	Forest (Coniferous)	26	3,960	10	47	43
Fall	Valley-Meadow	3	2,632	93	5	2
	Sagebrush	10	1,348	74	25	1
	Forest Park	6	1,265	56	29	15
	Herbland	3	165	45	55	0
	Forest (Coniferous)	9	2,929	52	8	40
		473	262,602			

¹ Data for valley areas involving 5,361 instances of use included from 44 sites examined by Martinka (1965).

² Winter, November 16-April 15; spring, April 16-June 15 in valley areas, May 16-June 30 in mountain areas; summer up to September 15; fall up to November 15.

³ May include grasslike sedge and *juncus* spp.

as the main food source for animals feeding within coniferous forest stands (Appendix VII). Combined use was 72 percent.

Spring

Grass continued to be the predominant class of forage used on most types over the spring season as a whole. Forb use about equaled or exceeded that on grass on half the types. The bunchgrass-shrub type was mainly used before forbs became readily available. Bluebunch wheatgrass and bluegrasses were the most important food items (Appendix V). Combined use averaged 58 percent.

Bluegrass and sedge species were the most important plants used on the valley meadow type (Appendix VI). Combined use was 55 percent. Bluegrass, Idaho fescue and balsamroot were the most important species used by elk moving onto or migrating over the outwash plain, sagebrush type during spring (Appendix VIII). Combined use on these species was 42 percent.

Sedge, mountain and nodding brome, geranium, dandelion and other forbs served as the main food items for elk feeding within the forest park type during spring (Appendix IX). Forb and grass use was 66 and 33 percent respectively. Mountain brome, sedge and geranium were the most important plants used on the burn type during spring (Appendix IX). Combined use was 58 percent. Mountain brome, slender wheatgrass, violet, and agoseris were the most important items on the high elevation herbland type. Combined use averaged 65 percent.

Bluegrass and a variety of forbs served as the most important food for elk feeding in deciduous forest stands (Appendix VII). Sedge and pinegrass were the most important individual items in coniferous stands. Aggregate forb and grass use was about equal in both deciduous and coniferous stands, but shrubs continued to be used to a greater extent within the latter.

Summer

Forbs were the predominant class of forage used by elk on all but the valley meadow and the highest elevation subalpine meadow types. Willows were the most important food item on the former type, averaging 70 percent (Appendix VI). Sedge and aster species and lupine were the most important food items on the subalpine meadow type, aggregating 62 percent (Appendix IX).

Lupine was the most important single food item on the sagebrush type over the summer as a whole, averaging 50 percent (Appendix VIII). It was used to the greatest extent after other vegetation cured during late summer or was killed by frost. Agoseris, buckwheat, little sunflower, and balsamroot were more important during the early summer. Combined use on these and lupine was 79 percent.

Agoseris, potentilla, dandelion, and aster were the most important forb species used by elk feeding on the forest park, burn and herbland types during the summer (Appendix IX). Combined use averaged

55, 42, and 60 percent for each type, respectively. These were obviously highly preferred forage plants which were selectively used even on sites where they were a minor component of the vegetation. The species had widespread distribution. However, they were relatively abundant on sites that had been disturbed by pocket gophers (Thomomys talpoides), on immature soils, and on ridgeline sites that elk may have maintained in a disclimax stage.

Aster, dandelion, and agoseris were the most important forbs used in both deciduous and coniferous forest stands during the summer (Appendix VII). These three species, nodding brome, wheat-grasses, and aspen received 51 percent of the elk use in deciduous stands. They received 56 percent, in combination with lupine, spirea, and huckleberry, in coniferous stands.

Fall

Grass was the predominant class of forage used on most types during fall. Bluegrass was the most important single item on sampled valley meadow types, averaging 91 percent of the recorded use (Appendix VI). Junegrass, bluebunch wheatgrass, bluegrass, and lupine were the most important on the sagebrush type. Combined use was 70 percent (Appendix VIII). The increased use of grass on these types appeared to be in response to most forbs losing their succulence and fall rains causing new leaf growth on bluegrasses and junegrass.

Mountain brome and lupine appeared to be the main plants used by the relatively few elk that continued to forage on the herbland type during fall (Appendix IX). Combined use was 59 percent. Sedges and the aster-fleabane-goldenrod group of plants were most important food items within the forest park type. Combined use was 76 percent. Geyer's sedge, pachystima, spirea, and huckleberry were the most important plants used by elk feeding within coniferous forest stands (Appendix VII). Combined use averaged 79 percent.

Seasonal and Yearlong Averages

The two forest types used during winter were estimated to provide at least 10 percent of the November 15 through April 15 diet for free-ranging elk as a whole. Proportionately weighted grass, forb, and shrub averages for the two forest types and the nonforest types which provided the main source of winter food were 69, 2, and 29 percent, respectively.

Respective grass, forb and shrub averages were 50, 45, and 5 percent for the different habitat types used during spring; 19, 62, and 19 percent for summer. The unadjusted values would be expected to vary with late springs resulting in an increased use of grass and shrubs; early springs, an increased use of forbs. The unweighted summer values probably would not vary greatly during years when animals used certain habitat types to a greater extent than others.

The general use of different habitat types through spring and summer was considered to preclude any meaningful weighting of averages.

Recorded observations of elk indicated that their relative use of forest and nonforest types may have been in a ratio of about 75 to 25 percent during the fall (Figure 11). By this relationship, fall use of grass, forbs, and shrubs averaged 56, 13, and 31 percent, respectively.

The winter, spring, summer, and fall forage class averages extended over about 5, 2, 3, and 2-month periods, respectively. Proportionate weighting of seasonal forage class figures gave an average yearlong food habit of about 51 percent grass, 26 percent forbs, and 23 percent shrubs for free-ranging elk. These values could be expected to vary with elk groups using particular vegetation types extensively, during years with extreme weather, or with animals that also used artificial food on feed grounds. What is perhaps best shown by the food habits study is that elk are extremely versatile and generalized feeders on all classes of forage and on a variety of plant species. Important winter food plants were major components in either seral stages or climax stands. This obviously enables the elk to contend with a broad spectrum of environmental change and persist as a faunal dominant in the Grand Teton and Yellowstone environments.

EFFECTS ON HABITATS

The effects of elk and other animals on habitats in valley and mountain areas were assessed from vegetation measurements, recorded notes at feeding sites, and general field observations.

Valley Areas

Measurements on shrubs and grasses were obtained from elk wintering areas within the bunchgrass-shrub type. Percent utilization of bunchgrasses was determined from permanent 100-plant sample units by the ungrazed plant method (Roach, 1950). Condition and leader use measurements on shrubs followed methods described by Cole (1963). Between 25 and 50 shrubs were sampled on each permanent unit. Measurements were made in May or June and additionally in October or November if domestic livestock grazing had occurred during the summer or fall. Data from 1967 were from Houston (1968a). Averages were rounded to the nearest five units.

Shrubs

Average leader use in shrub sample units in Grand Teton was 70, 35, 55, 55, 40, and 35 percent from 1962 through 1967, respectively (Table 17). Proportions of severely hedged plants averaged 80, 75, 30, 35, 10, and 10 percent over these same years. These values reflect the end effects of combined use by elk, comparatively smaller numbers

Table 17.--Average percent leader use and severely hedged serviceberry and bitterbrush plants on permanent 25-50 plant sample units located on elk wintering areas in Grand Teton, 1962-1967.

Year	Number Sample units	Percent Leader use	Percent Severely hedged
1962	6	70 (50-80)	80 (50-100)
1963	10	35 (10-90)	75 (15-95)
1964	11	55 (10-90)	30 (10-70)
1965	10	55 (20-90)	35 (5-85)
1966	10	40 (5-80)	10 (5-25)
1967	10	35 (10-85)	10 (5-30)

of mule deer (Odocoileus hemionus) and moose (Alces alces), and on some units, domestic cattle. Total use on shrubs was apparently greater prior to or during the severe 1961-62 winter. The 1962-63 winter could be classed as comparatively mild; 1963-64, average; 1964-65, severe; 1965-66 and 1966-67, average by snow depths alone (Table 1). The absence of crusting actually made the 1964-65 winter less severe than indicated.

Lighter leader use and progressive improvements in shrub conditions have apparently occurred over the series of less severe winters since 1961-62. Permanent sample units were not established on Douglas rabbitbrush because this shrub was in abundant supply and appeared able to withstand very heavy use.

Periodic summer inspections showed cattle began to forage on upland slopes of the bunchgrass-shrub type after vegetation on sagebrush flats started to cure. This occurred from the first week of August to the first week of October between 1963 and 1967. Utilization of bitterbrush by cattle on an area sampled by two units averaged 80, 40, 60, 30, and 5 percent from 1963 through 1967. On another area, their use on bitterbrush amounted to 50 percent of the available leaders in 1963; 5 percent in 1964, 1965, and 1966; none in 1967. Cattle use on serviceberry was usually lower than on bitterbrush, ranging from 5 to 10 percent on most sample units. Use up to 70

percent occurred on one unit which sampled an area which was grazed through October and early November of 1964. General reductions in livestock use of shrubs and grasses on these important wildlife winter ranges were achieved after 1964 by moving cattle to fenced, irrigated pastures before the vegetation on sagebrush flats fully cured.

Grass

Utilization of cured bluebunch wheatgrass on five Grand Teton sample units used by wildlife alone averaged 25 percent (10 to 40 percent) from 1963 to 1967 (Table 18). Yearly utilization on individual units ranged from 0 to 70 percent. Utilization on three other units which sampled wildlife wintering areas that were also used by cattle averaged 65 percent (45 to 75) over the same period. Utilization on individual units ranged from 15 to 80 percent. Cattle accounted for 40 to 85 percent of measured grass utilization.

Average utilization of cured bluebunch wheatgrass on the refuge between 1964 and 1967 was 70 percent (65 to 75) on three units which sampled areas within 1 mile of feed grounds, 50 percent (35 to 70) on four units between 1 and 3 miles distant, and 20 percent (15 to 40) on four units between 4 and 5 miles distant (Table 18). The 1963-64 through 1966-67 winters were not particularly severe and the utilization differences between sample units and years mainly resulted from differences in elk distributions or the numbers of animals freeranging off feed grounds.

Table 18.--Average percent utilization of bluebunch wheatgrass on 19 100-plant sample units on Grand Teton areas used by wildlife alone, sites additionally used by cattle and within distance zones from National Elk Refuge feedgrounds, 1963-1967.

No. of sample units	Location	Years					Remarks
		1963	1964	1965	1966	1967	
5	Grand Teton	20	25	10	25	40	Wildlife only
3	Grand Teton	45(?)	75(65)	65(30)	65(40)	75(30)	Wildlife and cattle ().
4	Refuge	--	15	15	20	40	4-5 miles
4	Refuge	--	55	35	35	70	1-3 miles
3	Refuge	--	75	65	70	70	Within 1 mile

The comparatively light (less than 30 percent) to moderate (30 to 55 percent) utilization of bluebunch wheatgrass by wildlife on sample units within Grand Teton and more than 1 mile from refuge feed grounds partially illustrate the abundant food sources available to the elk. Forage production of highly and moderately preferred elk food plants on a 56,300-acre portion of the winter range was calculated at 28.9 thousand dry weight tons in 1967 by Houston (1968a). After adjustments for allowable use on important forage plants, from 9 to 12 thousand tons were considered potentially usable as an elk food source. Elk were estimated to remove from 3 to 4.5 thousand tons of this food source in addition to consuming about 2.1 thousand tons of baled hay on feed grounds during the 1967-68 winter (Houston, 1969 writ. comm.).

Most Grand Teton and refuge sample units were on south, east, or west slopes where variable moderate (30 to 55 percent) to heavy (60 to 80 percent) utilization could be expected if the greater portion of the elk herd did not concentrate on feed grounds from January through March. This was partially illustrated from calculations by Houston (1968a) which showed average utilization on the eight sample units greater than 1 mile from refuge feed grounds increased from about 30 to 60 percent with only a 15 percent increase in elk freeranging between 1965-66 and 1966-67 winters.

Vegetation on some slope areas within 1 mile of feed grounds has probably received consistent heavy utilization for many years. Bluebunch wheatgrass and Douglas rabbitbrush plants continue to persist on coarse soils, though their vigor, density, and forage production is reduced from that on more distant sites. Vegetation on most sampled slope areas 1 mile or more from feed grounds has probably received mostly light to moderate, but occasionally heavy, utilization. Results from clipping studies by Houston (1968a) show that total grass and shrub production on slope areas within 1 mile of feed grounds averaged about 50 percent (25 to 70) of site potential; 1 to 3 miles, about 60 percent (41 to 78); 4 to 5 miles, about 75 percent (50 to 105). Bluebunch wheatgrass production averaged about 35, 50, and 70 percent of potential within these respective zones. Douglas rabbitbrush production on most sample areas exceeded that on site potential units and was indicated to increase with reduced grass production.

Needle-and-thread and Sandberg bluegrass occurred in mixtures with bluebunch wheatgrass and Douglas rabbitbrush on fine textured soils. Proportions of these appeared to increase as bluebunch wheatgrass decreased on heavily grazed sites. Such biotic disclimaxses were most prevalent within 1 mile of feed grounds, but also occurred on ridgetop and upper slope areas throughout the winter range. These conditions, as well as the reduced vigor and densities of bluebunch wheatgrass on coarse soils adjacent to feed grounds, may have resulted

from spring as well as winter grazing. An average of 55 percent (48 to 58) of the bluebunch wheatgrass plants on the sample units within 1 mile of feed grounds had their new spring growth grazed from 1964 to 1966. Comparative averages for units between 1 and 3 miles distant were 25 percent (15 to 32); 4 and 5 miles, 10 percent (8 to 12). An average of 20 percent (16 to 23) of the bluebunch wheatgrass plants on slope sample units within Grand Teton had their new growth grazed between 1963 and 1966. Plants which had been closely cropped during the winter seemed to be selectively regrazed for their new growth in spring.

Effects

Old homestead sites and pastures that had not been used by livestock for the past 10 to 30 years were still evident on elk wintering areas within Grand Teton Park and the refuge. Vegetation contrasts along old fence lines and terrace steps on slopes showed sites that had once been very heavily grazed. Present vegetation stands on these and comparisons with ranges still being used by both elk and livestock suggested that elk use alone had permitted general improvements in plant density and vigor on bottomland, swale, and most slope areas. Possible exceptions were ridgetops and adjoining upper slope sites that occurred throughout the winter range and upland slopes and flats adjacent to refuge feed grounds. The vegetation on these sites

appeared to be maintained as a biotic disclimax. This term was considered to apply to the arrested development of vegetation to its climatic potential as well as retrogressions from higher to lower successional stages.

Upper slopes were delineated as areas which were kept free of snow by wind action or first thaws. These varied from an approximate one-third of the top to base distance on some steep south exposures to being nonexistent on north exposures. Other major portions of different slope exposures which were usually snow covered through the winter are collectively called slope areas. These and swales were principally bunchgrass-shrub vegetation. The predominant vegetation on bottomland areas was the valley meadow type.

Records on habitat use, the food habits study, and notes taken at elk feeding sites showed that certain plant species on bottomland, swale, and slope areas, which were usually partially or completely covered by snow, were the main food source for most free-ranging elk from about mid-November through March. Forest parks or coniferous forest types with understories of sedge and shrubs appeared to be ecological equivalents to bottomland and swale areas for small groups of wintering elk. The duration and intensity of elk use on these different areas varied with snow depths, crust conditions, the extent to which the animals pawed feeding craters or used forage sources.

More productive areas, such as bottomlands and swales, appeared to be used in preference to drier slope areas for as long as possible.

Observations on the main refuge and Grand Teton winter ranges suggested that the animals that were not artificially fed would usually use bottomland and swale areas predominantly from mid-November through January. Depending upon previously mentioned conditions, east, west, north or south slope exposures would be variously used from February through March, or the animals might remain longer on, or return to, bottomland or swale areas. The development of crust conditions on one slope exposure caused shifts to others or into forest types when these were available. Changes in snow structure, crust conditions, or depths with February, March, or April thaws allowed animals to move over snow to new forage sources or re-use areas.

Vegetation on ridgetop and adjoining upper slope sites, which remained snowfree because of wind action or first thaws, appeared to be too limited in supply and too easily obtained to represent a food source that ultimately determined elk numbers. Such sites were estimated to make up from 4 to 8 percent of major winter range complexes and 10 to 20 percent of their total slope and ridgetop areas. Consistently heavy winter and/or spring use of ridgetop and upper slope vegetation with variable animal distributions and weather conditions suggested that free-ranging elk would continue to maintain biotic disclimaxes

on these areas with relatively low population numbers. Bergerud (1967) also did not consider the most accessible forage on exposed slopes to be a limiting factor for free-ranging caribou (Rangifer tarandus).

Under the winter conditions of the study area, free-ranging elk did not seem to be able to progressively deplete their main plant food sources on slope areas that were usually snow covered. This was partially due to protection afforded by snow, but other factors seemed equally important. These were: an apparent adaptation of native forage plants to contend with periodic heavy winter use; the elks' variable use of different slope exposures; the presence and alternative use of more abundant and preferred food sources on bottom-land and swale areas; and finally, death and reproductive rate adjustments in the elk population. It appeared that food sources on slope areas could be the most limited during certain winters, but this would depend upon the extent that other areas could not be used.

Free-ranging elk also seemed unable to progressively deplete their grass and grasslike plant food sources on bottomland areas, or the grasses and shrubs which represented their main winter food within swales. Inspections, which included the most heavily used bottomlands and swales adjoining feedgrounds, suggested that the overall utilization of herbaceous vegetation by elk on such areas

was limited to light or moderate levels by the necessity of pawing snow to obtain forage. This limit on utilization resulted from a hardening and spacing of pawed feeding craters comparable to that reported by Pruitt (1959). Shrubs, which had more favorable moisture relations in swales and were also protected by snow accumulations, appeared able to withstand periodic heavy use and recover in interim years.

Elk did consistently overuse some willow stands on bottomlands and contribute to their being replaced by other vegetation. The replacement of willow by grasses and grasslike plants and aspen by coniferous forest or herbaceous vegetation were successional processes which appeared to be mainly directed by climate and changing moisture or soil conditions for the different plant species. Elk participated in these processes by accelerating the replacement of seral willow and aspen in late succession stages or when stands were reduced to remnant status. These biotic effects, when palatable seral vegetation reached successional stages where it did not recover from periodic heavy use or was reduced to the extent that it no longer represented a food source that limited elk numbers, appeared inevitable. The rate of succession for late stages of seral vegetation could conceivably be slowed if elk numbers were held at levels which reduced the intensity of their effects, and if other biotic agents and/or invading or competing vegetation did not reestablish either the same or a more accelerated rate of change. More extensive stands of early successional

stages of willow and groves of uneven-aged aspen in deeper snow zones appeared to be able to recover from the periodic heavy use they received by elk and other wild ungulates. The biotic effects of elk on different successional stages of willow were similar to those reported for moose by Houston (1968b).

Mountain Areas

Vegetation measurements were obtained from elk summering areas within the herbland type. Sampled areas were inside and outside two 5-acre exclosures at about 9,000 feet and on other sites. The two exclosures were on ridgeline sites on Red Creek and Chicken Ridge areas in the central mountain region in southern Yellowstone Park. Exclosure sites had been selected to sample areas used by large numbers of summering elk. Equally high numbers used the north end of Big Game Ridge within Yellowstone Park. The Red Creek Ridge area received relatively greater June through July elk use and was most intensively studied.

Plant and Ground Cover

Percent plant canopy cover (Daubenmire, 1959) and bare ground were measured on four sample units on top of Red Creek Ridge. A fifth unit was measured inside the exclosure. Measurements were made in August of 1962. Appendix II shows a summary of results. Plant cover totals exceed 100 percent because of overlapping plant crowns.

The exclosure and YG1 units were on comparable ridgetop sites which were bordered by extensive slope herblands on the south, east, and west and adjoined swales of the subalpine meadow type on the north. The meadow and ridgetop sites were favored June-July bedding and loafing areas for groups of over 200 elk. The YG4 and 5 units sampled adjacent hermland slope areas that had been extensively disturbed by pocket gophers. On the YG4 unit, mounds or casts occurred within 23 of the 25 square yard plots and 13 were freshly dug; on the YG5 unit, 20 of the 25 plots and 14 were fresh. Plants had their canopy cover partially reduced by elk utilization. The measurements mainly reflect the predominance of forbs within the hermland type and illustrate the intensity of pocket gopher activities on some sites.

Grass Condition

Five sample units were established to measure the condition and relative densities (Cole, 1963) of Idaho fescue plants on ridgetop sites inside and outside the Red Creek exclosure (Table 19). Units Y1 and 2 and Y3 and 4 were paired. The two outside units were within about 50 feet of the exclosure fence. Measurements showed the effects of elk grazing in maintaining Idaho fescue plants in a fair to poor condition. Densities were not appreciably different. Leaf height differences on outside units reflected current use. Comparable conditions occurred on limited sites at low elevations. These were on areas adjoining natural licks and limited ridge and slope areas that bordered forest parks which were used by large numbers of elk in spring.

Table 19.--Condition, relative density, and maximum leaf heights of Idaho fescue plants inside and outside an exclosure on Red Creek Ridge.

Plot No.:	Y2	Y1	Y4	Y3	Y5
Location:	Exclosure	Outside	Exclosure	Outside	Outside
No. plants:	50	50	50	50	100
Percent HC and CE ¹	30	54	40	66	32
Condition class ²	Good	Fair	Good	Poor	Good
Density index ³	14.7	12.8	11.5	12.6	5.5
Maximum leaf height (inches)	4.0	2.9	4.2	2.4	2.4

¹ Hollow center (HC) and clump edge (CE) plants.

² Rating from percent of hollow center and clump edge plants in sample.

³ Average distance in inches between sample point and closest plant.

Unit Y5 was located about 150 feet from the Red Creek enclosure on the next closest comparable ridgetop site. The measurements suggested that the principal effects of elk on Idaho fescue plants away from the immediate vicinity of bedding sites and the enclosure was to reduce leaf heights. Cropped portions of plants indicated at least 47 percent had been spring grazed.

Use of Key Plant Species

Grazed plants in 100-plant samples (of each) of two key grasses and one forb were tallied to measure the intensities of spring and summer elk use on herbland sites. Mountain brome, slender wheatgrass, and agoseris were selected on the basis of food habits studies. Grass plants were tallied as spring grazed if cropped leaf ends showed they had been used during early growth periods, as summer grazed if evenly cropped leaves or culms showed they had been used at a mature stage. Forbs were considered summer grazed if any portion of the plant was used. Tallies were made during late August and September when vegetation was mature and larger elk groups had dispersed from subalpine areas. Results from permanently established units that sampled slope areas south of the Red Creek enclosure are shown in Table 20.

Spring use on the key grasses varied between years. Summer use was uniformly low. Moderate intensities of summer use occurred on agoseris. Sample units and field observations on other subalpine

Table 20.--Percentages of key food plants grazed during spring and summer periods, 1964-1967.

Years and use period	Slender Wheatgrass	Mountain Brome	Agoseris
1964			
Spring	58	50	
Summer	8	2	31
1965			
Spring	37	25	
Summer	12	4	36
1966			
Spring	27	16	
Summer	11	2	31
1967¹			
Spring	45	44	
Summer	14	6	40

¹ From Houston (1968a).

areas showed these same plants were less heavily used than on the Red Creek sites. Summer use of 100-plant samples of mountain brome, slender wheatgrass, and agoseris on the most heavily used portions of Chicken Ridge in 1964 was 19, 10, and 34 percent, respectively. Two agoseris units on the north portions of Big Game Ridge had 6 and 9 percent use this same year.

The relatively heavy spring use of mountain brome and slender wheatgrass in 1964 provided an opportunity to assess the short term effects of such use. September measurements of 150 plants showed the average height of spring grazed mountain brome plants was 17.5 inches as compared to 19 inches for ungrazed plants. Spring grazed heights for slender wheatgrass on the sample unit averaged 13.5 inches in comparison with 15 inches for ungrazed plants. All ungrazed plants produced seedheads. About 2 percent of the spring grazed mountain brome and 8 percent of the slender wheatgrass plants did not produce seeds.

Vegetation Trends

Eleven "loop frequency" range transects (Parker, 1951) that were established in 1959 were remeasured in 1965. Transects were inside and outside Red Creek and Chicken Ridge exclosures. Sampled sites outside exclosures received relatively greater elk use than subalpine areas as a whole, because the transects were adjacent to ridgetop bedding areas

and/or close to exclosure fences. Comparatively light use occurred within exclosures by some elk entering on snowbanks or breaking down fences. Results from comparing loop hits on vegetation are shown in Table 21. Comparable transects are grouped. Hypotheses of similarity between 1959 and 1965 measurements could not be rejected for ridgeline transects 1 and 2E. Chi-square values led to rejections of similarity at the 0.2 level for S5 and S11 measurements; at the 0.1 level for RT9; at the 0.05 or 0.01 levels for all others.

Effects

Vegetation on the heavily used Red Creek ridge site sampled by the RT1 transect has not changed significantly since 1959, but slight increases in perennial forbs may have occurred. Plant densities on RT2E may have been close to site potential in 1959 and large changes did not occur under protection from elk grazing. Transect RT3E sampled a swale ecotone with herbland which was probably extensively disturbed by pocket gophers in 1959 (61 bare ground hits). Grasses and forbs significantly increased under protection from elk and an apparent reduction in gopher activity.

Herbland slope transects 4, 5, 6E, and 7E apparently had a sparser perennial plant cover in 1959. Total perennial vegetation had increased on all sites by 1965, but forbs to a greater degree within the exclosure areas. Ellison (1954) reported that selective

Table 21.--Comparison between 1959 and 1965 100 3/4-inch loop hits on 11 transects inside and outside exclosures on high elevation elk summer range areas in Yellowstone National Park.

Transect No. ¹	Perennial forb hits				Other hits ²		1959-65 trend	
	Grass 1959	hits 1965	1959	1965	1959	1965	Grass	Perennial forbs
RT1	12.0	13.5	45.0	55.5	43(34)	31(21)	+1.5	+10.5
RT2E	32.0	30.5	39.0	48.5	29(24)	21(7)	-1.5	+9.5
RT3E	2.0	23.5	22.0	34.5	76(61)	42(21)	+21.5	+12.5
S4	8.5	25.0	28.5	31.0	63(52)	44(22)	+16.5	+2.5
S5	8.5	17.0	31.5	35.5	60(45)	47.5(33)	+8.5	+4.0
S6E	8.5	13.5	23.0	40.0	68.5(45)	46.5(28)	+5.0	+17.0
S7E	6.0	12.0	21.0	55.0	73(55)	33(14)	+6.0	+34.0
RT8	15.0	23.0	12.0	20.5	73(40)	56.5(12)	+8.0	+8.5
RT9E	13.5	26.0	15.5	9.5	71(42)	64.5(30)	+12.5	-6.0
S10	9.0	22.5	16.0	19.0	75(39)	58.5(36)	+13.5	+3.0
S11E	16.0	9.0	13.0	21.0	71(32)	70(40)	-7.0	+8.0

¹ RT - ridgетop; S - slope; E - exclosure. Comparable transects - - -

² Litter, rock and bare ground hits with latter shown in ().

heavy grazing of forbs by domestic sheep in the hermland type favored increases in certain grass species.

Comparisons between the ridgeline and slope transects on Chicken Ridge were complicated by extensive pocket gopher digging within the enclosure. Gopher digging outside the enclosure appeared slightly less extensive. Both grass and perennial forb densities increased on the ridgeline transect used by elk between 1959 and 1965. Perennial forbs may have slightly decreased on the RT9E transect as a result of gopher activity within the enclosure. As on the Red Creek hermland, grasses increased while perennial forbs showed little change on the S10 slope transect. Grass may have decreased as a result of gopher activity on sites sampled by the S11 enclosure transect.

More bare ground was present when the Red Creek transects were established in 1959. Bendt (1960) reported pocket gopher numbers were at high levels. Excluding the near site potential transect, bare ground and annual forb hits on ridgeline sites decreased from an average of 52 hits in 1959 to 22 in 1965; on slope sites, 50 to 28 hits. On the gopher disturbed Chicken Ridge transects, bare ground and annual forb hits increased from an average of 46 to 53 hits on ridgeline sites and 42 to 56 on slope sites. The comparisons suggest that improvements (by reductions in bare ground or annual forbs) have apparently occurred with reduced gopher activity on Red Creek ridge sites. By these same criteria, downward trends occurred from combined elk and pocket gopher, or gopher use alone, on Chicken Ridge.

Photographs of ridgeline transect areas in 1959 show the vegetation had the appearance of a low canopy of perennial forbs. This is illustrated by the RT1 photo (Figure 12). Photos of slopes, as illustrated by the S4 transect (Figure 13), had the typical aspect of the herbland type. However, closeups of the different transect areas showed pocket gopher mounds were prevalent in the interspaces between larger plants. The appearance of ridgeline rather than herbland vegetation may have contributed to the general reports of subalpine ranges being in a deteriorated condition in 1959. Persons who constructed exclosures or visited sites reported that, in addition to the indicated heavy elk use and pocket gopher activity, Mormon crickets (Locustidae) and grasshoppers (Arididae) were abundant within the short vegetation on ridgeline areas.

Interpretations of elk effects on individual species could not be realistically made because the 100 3/4-inch loop hits on each transect did not give adequate samples. Differences between total hits on groups of species that were preferred spring and/or summer foods were examined as a possible indicator of relationships.

From 1959 to 1965, dandelion, agoseris, potentilla, aster, and fleabane species increased a total of 14.5 hits on the RT1 transect. Comparable increases on the RT2 and RT3 transects within the Red Creek exclosure were 12.5 and 19.0 hits. Totals for potentilla, aster,

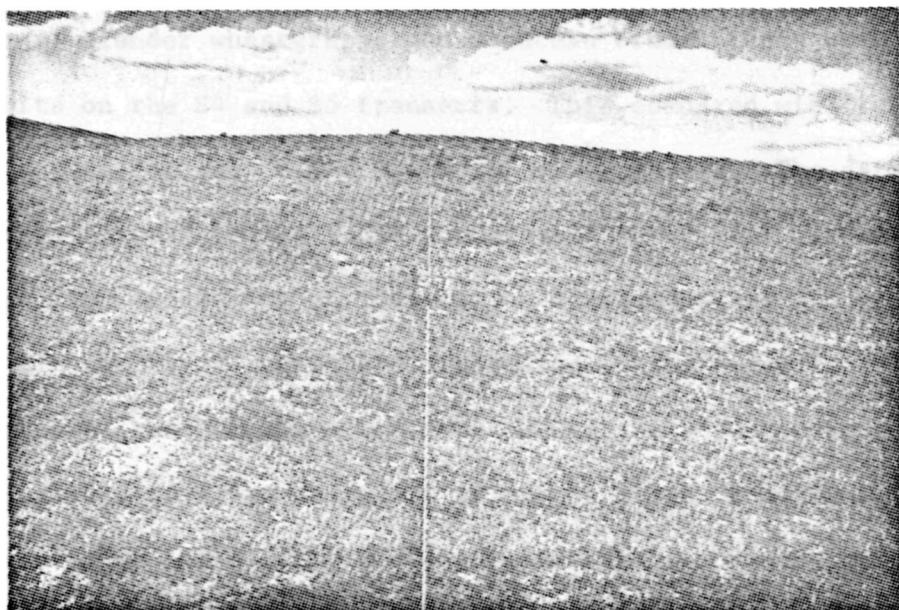


Fig. 12. Ridgetop photo.

plant growing conditions, pocket gopher activity, and the
soil on some hillsides where were highly
eroded.



Fig. 13. Herbland photo.

fleabane, slender wheatgrass, and mountain brome plants were 16.5 and 12.5 hits on the S4 and S5 transects. This compared with totals of 12 hits each for the S6 and S7 exclosure transects. The data from Red Creek suggest that elk did not prevent their more preferred forages from increasing on ridgetop or herbland sites and protection alone did not favor substantial increases in preferred food species. By contrast, the Chicken Ridge data showed the pocket gopher, which appeared able to cause decreases in either grass or perennial forbs, may be a major contributor to vegetation changes on subalpine ranges. The apparent recovery of ridgetop and slope vegetation on Red Creek after the gopher population high in 1959 would suggest that vegetation changes by gophers, with complementing influences from elk, Mormon crickets, and grasshoppers, were temporary.

Plant growing conditions, pocket gopher activity, and intensities of elk use on subalpine ranges were highly variable from 1962 through 1966. The measurements on or adjacent to elk bedding or loafing sites and close to exclosures show the animals maintained what could be considered biotic disclimaxes. Effects over subalpine ranges as a whole appeared to be limited to comparatively light use of current forage growth during the study period.

It seems certain that particular combinations of biotic and weather influences will periodically cause these subalpine ranges

to appear in what could be interpreted as a deteriorated condition. These interpretations should be tempered with ecological considerations. Subalpine vegetation has evolved over long periods of time by natural selection processes. Selection pressures would have had to result in vegetation having the capacity to contend with the periodically harsh subalpine environment and native animal life. Since the subalpine areas studied were not being subjected to influences that could be considered unnatural, it seemed unlikely that their variable appearance or temporary vegetation changes represented a departure from natural relationships.

Other Areas

The effects of elk on vegetation and the general condition of subalpine ranges on the portions of Big Game Ridge south of Yellowstone Park have been variously interpreted. Beetle (1952) considered this area to be in "the poorest condition class." This classification may have partially resulted from interpretations that herbland vegetation should be predominantly grasses or sedges (Beetle, 1962). Anderson (1958) considered gullied upper slopes to be badly depleted, but recognized the presence of unstable soils. Croft compiled a report from an August 18-23 inspection trip in 1955 (Croft and Ellison, 1960). He concluded ". . . the destruction of vegetation and soil . . . is largely a result of excessively heavy use by elk." The report also suggest elk were responsible for flood damage and mud-rock flows.

The northern portion of Big Game Ridge inside Yellowstone Park was routinely covered during this study. Southern portions could be seen from the top of Chicken Ridge. These were visited for comparison purposes during late summer or fall periods. Aerial flights also sampled both areas. The subalpine portions of northern Big Game Ridge may have escaped an 1885 fire. Upper slope areas were covered by the most dense herbland stands encountered on the study area. Summer elk densities were higher, in relation to the amounts of herbland, than on other ridge areas.

Upper slope areas on the southern portion of Big Game Ridge were a series of sparsely vegetated drainage head gullies and slopes. Soils were derived from residual sandstone and shale formations. These drainage head conditions may have resulted from the 1885 or earlier fires removing dense herbland vegetation (comparable to that on northern portions) and later high intensity fall or spring rains causing gullying and soil losses. They may also exist primarily as a result of unstable or sandy soils occurring in a harsh drainage head situation and not have been caused by fire.

A distinct pioneer or post-fire type of perennial vegetation does not appear to have evolved in the subalpine zone. Annual forbs may serve as ecological equivalents until perennial vegetation again reoccupies sites. The establishment of dense stands of perennial

vegetation on porous sandy soils may not be possible or may require an extremely long period of time on harsh ridgetop and upper slope sites at high elevations (9,500 to 9,800 feet). Actively moving clay soils will probably eventually stabilize and support a denser vegetation cover.

Drainage head areas on south Big Game Ridge and other portions of the study area commonly retained snowbanks into July. Vegetation within these was normally absent or very sparse. Such areas were snowfree in August and could be misinterpreted as illustrating elk effects. Scattered upper slope areas with stable soils and lower slopes had stands of subalpine vegetation comparable to the northern portions of Big Game Ridge and other Yellowstone areas.

ELK MANAGEMENT

Regulations

Grand Teton is the only national park where its enabling legislation allows a native wild animal to be killed by state-licensed hunters deputized as rangers. This is provided for under Public Law 787 of the 81st Congress. The provision applies only to elk and to portions of lands added to the original Grand Teton National Park in 1950. The pertinent section of Public Law 787 which relates to the elk is as follows:

"Sec. 6 (a) The Wyoming Game and Fish Commission and the National Park Service shall devise, from technical information and other pertinent data assembled or produced by necessary field studies or investigations conducted jointly by the technical and administrative personnel of the agencies involved, and recommend to the Secretary of the Interior and the Governor of Wyoming for their joint approval, a program to insure the permanent conservation of the elk within the Grand Teton National Park established by this Act. Such program shall include the controlled reduction of elk in such park, by hunters licensed by the State of Wyoming and deputized as rangers by the Secretary of the Interior, when it is found necessary for the purpose of proper management and protection of the elk."

The use of licensed hunters as deputy rangers appears to circumvent an International Treaty for "Nature Protection and Wildlife Preservation in the Western Hemisphere" between the United States of America and 17 other countries (National Park Service, 1968). These contracting governments agreed to "prohibit hunting, killing, and capturing of the fauna and destruction or collection of representatives of the flora in national parks except by or under the direction or control of the park authorities, or for duly authorized scientific investigations."

A permit is required to kill an elk within Grand Teton National Park. The number of permits to be issued each year is determined jointly from studies by Park Service and Wyoming biologists. Permits are requested through the Wyoming Fish and Game Commission's office which furnishes the Secretary of the Interior with a list of licensed hunters. Permits are issued to hunters authorized as deputy rangers at a check station within Grand Teton Park.

Park hunt units during 1966 are shown on Figure 14. Unit 4 was a roadless mountain area. Roads provided easy access to and through the other three hunt units. Unit 3 was largely foothill terrain with coniferous forests and interspersed parklands. Units 1 and 2 were established as hunt units in 1963. Unit 2 included the eastern side of the forested Snake River bottom, the adjoining open sagebrush-covered benches and during 1965 and 1966, the northern half of

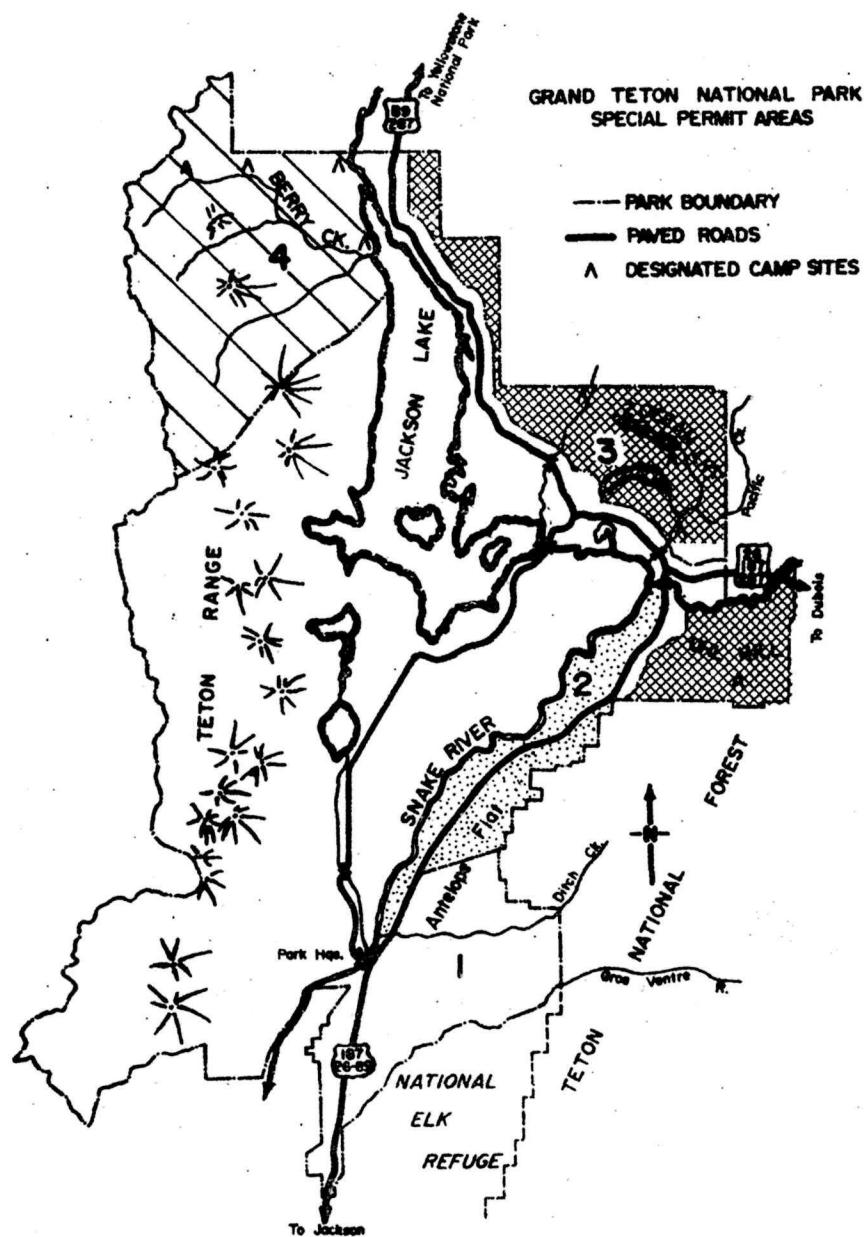


Fig. 14. Map of park hunt units.

Antelope Flat, as shown. Unit 1 included the partially forested Blacktail Butte, as well as extensive sagebrush areas, hayfields, and pasture lands on the southern half of Antelope Flat during 1965 and 1966 seasons. In 1963 and 1964, it also included the northern portion of Antelope Flat.

Permits were valid for entire hunting seasons before 1963. In 1963, 1,000 A permits were valid for an October 1 through November 15 period; 1,000 B permits for an October 15 through November 15 period. In 1964, 500 A, 1,000 B, and 1,500 C permits became valid on October 1, 15, and November 1, respectively. The season closed on November 30. In 1965, 1,000 A and 2,000 B permits became valid on October 9 and 30, respectively, and the season ran to December 15. In 1966, 1,000 A and 1,500 B permits became valid on October 15 and November 5. This season closed November 30.

All holders of valid permits could hunt in units 3 and 4. Numbers of hunters in units 1 and 2 were controlled by issuing 50 special permits for each unit by weekly periods or within weekly periods. Special permits were issued to regular permit holders on a first-come, first-served basis, or by a drawing if applicants exceeded available permits. Areas 1/4 mile on either side of main highways and 1/2 mile from buildings were closed to hunting.

Purpose of Management

Public Law 787 specifically restricts the hunting of elk within Grand Teton to " . . . when it is found necessary for the purpose of proper management and protection of the elk." This "hunting when necessary" restriction was variously interpreted. Reviews of early records and the results from 1962 and 1963 field studies resulted in management problems being defined as follows:

1. A large herd of elk which summered in Grand Teton Park was migrating to the refuge in October and using forage which should have been reserved for winter periods.
2. Late migrating herd segments from southern Yellowstone Park, which traveled through roadless wilderness areas before they crossed Grand Teton, had become too large to manage without assistance from park hunts.
3. Herd segments that migrated through roaded areas east of Grand Teton or summered on the more accessible national forest lands between the two parks had been reduced to levels where they no longer represented the major portion of the elk herd.

This definition of problems led to cooperative management programs which had the long-term objectives of restoring historical elk distributions and migration patterns and reducing the need to hunt elk within Grand Teton Park. Yearly programs after 1963 attempted to

(1) restrict early migrations of Grand Teton elk, (2) progressively reduce late migrating Yellowstone herd segments to levels where they could eventually be managed by hunting outside Grand Teton, and (3) allow compensating increases in the herd segments that migrated east of Grand Teton or summered on national forest lands between the two parks.

Hunt Statistics

Yearly and Total Elk Kills by Units

Table 22 shows legal kills of elk in park units open to hunting between 1951 and 1966. Hunts did not occur on park lands in 1959 and 1960. Only units 3 and 4 were open to hunting up to 1962.

Yearly kills averaged 189 (27 to 325) animals. The additional use of units 1 and 2 between 1963 and 1966 contributed to increasing the average kill to 670 (612 to 753) animals. Proportion calculations that related known park kills to checked park and other area kills at a State checking station showed park kills ranged from 48 to 73 percent of total kills from the refuge herd during this period.

These figures undoubtedly represent increases over previous years.

Yearly kills in unit 4, which was a 49,000 acre roadless mountain area, were comparatively insignificant from a herd management viewpoint. The lower kills in units 1 and 2 in 1966 resulted from an early closing of these areas after a desired kill was obtained from the

Table 22.--Legal kills of elk in Grand Teton hunting units,
1951-1966¹.

	Kill by hunt units					Total kills	Percent of refuge herd kill
	4	3	2	1	Unknown		
1951-62 (avg.)	15 ²	182	---	---	3	189	Avg.
1963	23	192	164	246	0	625	48
1964	17	170	163	398	5	753	61
1965	9	208	155	319	0	691	73
1966	5	434	70	103	0	612	64
1951-66 Totals	85	2,837	552	1,066	8	4,548	

¹ Park closed to hunting in 1959 and 1960.

² Season only open in 1952 and 1962 and a 2-year average.

Grand Teton summer herd. This early closure encouraged hunting in unit 3 where increased kills were obtained from southern Yellowstone Park elk.

Permit Use, Season Dates, and Hunting Success

Applications for Grand Teton elk permits were taken as received until set numbers were reached. Averages from 1961 through 1966 show about 10 percent of the permits were obtained by local residents from within the Jackson Hole area. Other Wyoming residents obtained 80 percent of the permits and nonresidents, 10 percent.

Table 23 lists the dates, numbers of permits authorized and used and hunter success rates for park seasons since 1951. This information shows season starting dates have varied from September 10 to October 20; closing dates, from October 31 to December 15. As will be shown later, the coinciding presence of migrating elk and hunters, rather than season dates themselves, determined total elk kills and hunting success.

Before 1963, an average of 50 percent of the authorized 1,200 to 2,000 yearly permits was used. About 25 percent of the persons using their permits killed an elk. Since 1963, an average of 60 percent of the 2,000 to 3,000 yearly permits was used. Hunting success averaged about 51 percent for 1963 and 1964 and about 37 percent for 1965 and 1966. The lower success after 1964 appeared to result mainly from previously unhunted (since 1951) Grand Teton elk curtailing their early migrations or altering their fall distributions to stay within

Table 23.--Season dates, numbers of permits authorized and used, and percent hunter success, 1951-1966.

Year	Season dates	Permits authorized	Permits utilized		No. elk killed	Percent hunter success	Percent kill/total permits
			No.	Pct.			
1951	Sept. 10-Oct. 31	1,200	510	42	184	36	15
1952	Sept. 10-Nov. 16	1,200	455	38	27	6	2
1953	Sept. 10-Nov. 5	1,200	568	47	112	20	9
1954	Sept. 10-Dec. 12	1,200	600	50	104	17	8
1955	Oct. 20-Nov. 20	1,200	624	52	310	50	26
1957	Oct. 20-Dec. 10	1,200	748	62	160	21	13
1958	Oct. 20-Nov. 30	1,200	583	48	110	19	9
1961	Oct. 15-Nov. 30	2,000	1,002	50	278	28	14
1962	Oct. 1-Dec. 2	2,000	1,170	58	280	24	14
1963	Oct. 1-Nov. 15	2,000	1,194	60	625	52	31
1964	Oct. 1-Nov. 30	3,000	1,506	50	753	50	25
1965	Oct. 9-Dec. 15	3,000	1,943	65	691	36	23
1966	Oct. 15-Nov. 30	2,500	1,615	65	612	38	24
Totals		24,100	13,294	55	4,570	34	19

closed areas west of the Snake River. Some groups of particularly vulnerable resident elk that summered along the east side of the Snake River in what became unit 2 may also have been reduced by the 1963 and 1964 seasons.

About 72 percent of the 1,000 A and 47 percent of the 1,000 B permits were used by authorized park hunters in 1963. In 1964, about 72, 62, and 44 percent of the 500 A, 1,000 B and 1,500 C permits were used. In 1965, about 69 percent of 1,000 A and 62 percent of 2,000 B permits were used. About 72 percent of 1,000 A and 60 percent of 1,500 B permits were used in 1966. This split permit system was used along with information sheets on elk migration habits to encourage hunters to hunt when elk were present. The A permits which allowed hunting over the entire season were most desired and used by hunters. The B permits, which became valid about 15 to 20 days after a season started, were fairly successful as a management tool after 1963. The C permits used in 1964 appeared to represent a management refinement that could not be used with a limited supply of hunters. These permits were not even fully applied for, despite the use of news and radio media.

Questionnaires were mailed in 1962, 1963, and 1964 to determine the reasons why park permits were not used. Replies for 1962 and 1963 were almost identical. Averages showed about 9 percent killed an elk elsewhere before their permit became valid; 21 percent killed an elk

elsewhere after their permit became valid; 32 percent hunted elsewhere; 38 percent did not hunt at all. Replies for 1964, when C permits were used, showed 30 percent killed elk elsewhere before their permit became valid; 8 percent killed elk elsewhere after their permit became valid; 27 percent hunted elsewhere; 35 percent did not hunt. The reasons that park permits were not used could be summarized as: approximately two-thirds of these permit holders hunted elsewhere and one-third did not hunt at all. Restorations of elk herds throughout Wyoming have given resident hunters many alternative places to hunt much closer to the State's population centers.

Composition of Kill

Complete or almost complete yearly samples showed averages of 45 percent females, 17 percent calves, 11 percent yearling males, and 27 percent adult males were taken during park hunts from 1955 through 1966. An average of 75 males was killed for each 100 females before 1963; 104 males per 100 females since.

Yearly samples of 568 to 721 elk were aged between 1963 and 1966. Averages of 32, 64, and 4 percent of the male elk older than calves were yearlings, 2-to-7-year-olds, and 8-year-and-older animals, respectively. Comparable figures for females were 15, 66, and 19 percent. The data show hunter selection contributed to the male segment of the population having a younger age structure than the female.

About 6 percent of the male elk older than yearlings were 8 years-and-older, as compared to 22 percent of the females. Percentages of 8-year-and-older elk in samples also suggested that sustained high hunting kills after 1963 reduced the proportion of older animals in the elk herd. Eight-year-and-older male elk comprised about 7, 2, 3, and 2 percent of the kill of males older than calves from 1964 through 1967. Comparable figures for female elk 8-years-and-older were 24, 22, 19, and 13 percent over the same series of years.

Evaluation of Management Program

Hunt System

The system of adjusting season dates to coincide with elk migrations, allocating general permits for two or three hunt periods and special permits by weeks for two additional hunt units since 1963 has, in comparison with 1962 (and previous years), greatly increased the efficiency of park elk management programs.

An average of 4 elk per day were killed during the October 1 - December 2 season in 1962; 14 per day during the October 1 - November 15 split A and B permit season in 1963; 12 per day during the October 1 - November 30 split, A, B, and C permit season in 1964; 10 per day during the October 9 - December 15 A and B permit season in 1965; 13 per day during the October 15 - November 30 split permit season in 1966.

Over 75 percent of 280 elk killed during the 2,000 permit 1962 season were taken during October. Large numbers of elk migrated through park hunt units during November, but insufficient numbers of hunters were present to secure significant kills. From 1963 through 1966, split permit periods, later season starting dates and manipulations of hunter numbers in two special hunt units resulted in November kills increasing from 50 to 75 percent of total kills. Total kills ranged from 612 to 753 animals and, in combination with kills outside the park, gave removals that maintained approximately stable herd numbers over the period.

Efficiency of Hunt Units

An average of 13.5 elk (5 to 23) per year was killed in unit 4 over the 4 years between 1963 and 1966 (Table 24). Comparatively few permit holders (20 to 50) appeared to have the desire or equipment to hunt elk in this 49,000-acre roadless wilderness area. The average success rate of 40 percent (25 to 57) was only slightly above that for units 2 and 3 and below unit 1. About 40 percent of the 54 persons killing elk over the 4 years were general public hunters who owned their own pack and saddle horses, about 25 percent were clients of commercial outfitters, and 35 percent were State employees, or guests of employees, who were involved in law enforcement activities within and outside the park. These records, plus results from field studies which showed hunting was not effective in causing movements into

Table 24.--Elk kills, numbers of hunters, and percent success in four park hunting units, 1963-1966.

Unit	Elk kills								Numbers of hunters				Percent success			
	1963		1964		1965		1966		1963	1964	1965	1966	1963	1964	1965	1966
	No.	Pct.	No.	Pct.	No.	Pct.	No.	Pct.								
1	246	39	398	53	319	46	103	17	350	550	630	150	70	72	51	69
2	164	26	163	22	155	23	70	11	350	550	553	150	47	30	28	47
3	192	31	170	23	208	30	434	71	444	376	731	1295	43	45	28	33
4	23	4	17	2	9	1	5	1	50	30	29	20	46	57	31	25
Totals	625		748 ¹		691		612		1194	1506	1943	1615	52	50	36	38

¹ Area unknown for 5 kills.

adjoining hunt areas (Cole, 1967), indicated that the contribution of unit 4 to the overall management program was small.

An average of 251 elk (170 to 434) per year was killed in unit 3 over the 1963 to 1966 period. Highest kills were made in 1965 and 1966 when units 1 and 2 were closed during mid-season to direct hunting pressure onto large numbers of migratory elk within the unit. Numbers of hunters ranged from 376 to 1,295, with lowest use during years when the greatest numbers of permits were available in units 1 and 2. The average success rate was 37 percent (28 to 45). This hunt unit obtained kills from the relatively large late migrating segment from southern Yellowstone Park. Plowing snow from roads permitted late season hunting in this unit and adjoining areas outside the park.

Unit 3 appears essential to the management program until the late migrating Yellowstone elk can be controlled by hunting outside park boundaries. This could be accomplished by strategically locating public parking and camping facilities between Grand Teton and the south boundaries of the Teton Wilderness Area, by developing foot and horse trails that bisect elk migration routes, and plowing main access roads as needed.

An average of 138 elk (70 to 164) per year was killed in unit 2. Hunting success for the 150 to 550 hunters using the unit averaged 38 percent (28 to 47). Before 1965, hunters were less interested

in this special unit which had only the main highway as its east boundary and dead end access roads leading to river bottom forests along the east side of the Snake River. Adding the north half of Antelope Flat (from unit 1) on the east side of the main valley highway to unit 2 reduced differences in applications for permits and hunting success between the two special units.

An average of 267 elk (103 to 398) per year was killed in unit 1. Hunting success in this extensively roaded unit averaged 65 percent (51 to 72). Numbers of hunters ranged from 150 to 630. The attraction of unit 1, and unit 2 after 1965, appeared to be the opportunity to drive an extensive main and secondary road system, see large groups of migrating elk, and kill an animal in close proximity to a road. Both special units were crossed by Grand Teton elk and eventually by the greater portion of the animals that also migrated through unit 3.

Units 1 and 2, in combination with similar permit hunts on the refuge, have been successfully used to curtail early migrations and reserve winter forage on the refuge; control the size of the Grand Teton summer herd; and during 1963 and 1964, assist in securing greater kills from late migrating elk. With the exception of the portion of unit 2 west of the main through highway, these two special hunt units appear essential to future programs. They should continue to be rigidly controlled by the limited special permit system and only used when desired elk kills cannot be obtained from other units or off park lands.

The use of units 1 and 2 to control the size of Grand Teton's summer herd could be reduced to the extent that additional elk kills were obtained from refuge hunts. The usual combined refuge and special park unit kill quota for this herd was about 400 elk. Only 20 permits were issued for weekly hunting periods on the refuge during 1963 and 1964; 40 permits, during 1965 and 1966. An average of about 90 elk (48 to 133) was killed annually within the 22,700-acre area. This compares with a 405 average for the two adjoining 19,000 and 12,500-acre special hunt units within Grand Teton Park.

Refuge and adjoining Forest Service lands comprised over 37,000 acres. The greater isolation of these lands from main roads and their interspersed tree cover and rough terrain make them a more esthetically suitable elk hunting area than the two special park units which were mainly open sagebrush flats. The use of refuge and adjoining Forest Service lands as hunt units to discourage large groups of early migrating elk from using the southern portions of the winter range 3 to 6 weeks before other herd segments arrive appears essential to future management programs. Separations into north and south hunt units would be desirable. Earlier closing of a north hunt unit (usually the first to second week of November) would allow undisturbed elk to use abundant food sources on their historical wintering areas within Grand Teton and the northern half of the refuge. This would

reserve bottomland food sources along Flat Creek for December and later winter foraging. Such patterns of use were reported to occur with early elk herds (Preble, 1911).

Illegal Kills

Known illegal kills of elk and other wildlife in Grand Teton areas open to hunting over a 10-year 1957-66 period totaled 64 elk, 24 moose, 1 mule deer, and 1 antelope (*Antilocapra americana*). Totals of 118 elk, 5 moose, 3 mule deer, and 2 coyotes (*Canis latrans*) were known to be killed in park areas closed to hunting during the same period.

Arrest records show not all illegal kills were made by holders of Grand Teton elk permits, but comparisons between years indicated total illegal kills increased in open hunting areas as numbers of permit holders increased. Total illegal kills in closed areas tended to decrease after 1963. This was probably due to intensified law enforcement and previously closed areas, where illegal kills had been high, being opened to hunting.

Conflicts

The illegal kill record for Grand Teton is probably no worse than outside areas where public hunting occurs. It does, however, show that hunting for elk contributed to other park wildlife being killed and violations of closed areas. The illegal kills of elk and

other wildlife in closed areas represented the most serious conflict with the primary purpose of Grand Teton National Park which was, in part, to provide visitors with the opportunity to see and photograph native wildlife. Disturbances from illegal hunting or killing of animals commonly caused large groups of elk to leave areas where they were providing outstanding viewing and photography opportunities for park visitors. Roadside moose in areas both open and closed to hunting were either mistakenly shot for elk or for other unexplainable reasons. Actions taken to reduce illegal kills involved distributing informational literature to hunters, conspicuously posting open and closed hunting areas, increasing ranger patrols, and requiring that guns transported through the park be unloaded, cased or broken down.

Comments and letters showed some fall visitors to the park found hunters carrying guns, animals being shot, or other hunting activities on the eastern and northern portions of Grand Teton objectionable. Attempts to reduce this conflict involved prohibiting hunting 1/4 mile on either side of main highways, restricting hunter camps to specified areas apart from visitor campgrounds, using hunt units 1 and 2 (where hunting activities were most visible) only when necessary, and progressively changing the starting dates for park hunts from October 1 in 1964 to October 21 by 1967.

Overall Results

The cooperative management program between Wyoming and the National Park Service has involved coordinating field studies and law enforcement activities, exchanging study information and preparing joint reports, making combined recommendations as called for under Public Law 787, and, after 1963, cooperatively monitoring the overall program designed to restore historical elk distributions and migrations.

The monitoring system used known total park kills and kills checked at a State hunter checking station to calculate the numbers of elk killed in seven different Forest Service areas. These, in combination with park kill figures and migration records, were used to measure in-season progress toward achieving set kill quotas and calculate removals from different migratory segments.

The management objective of discouraging early migrating Grand Teton elk from utilizing winter food on the refuge during October was largely achieved after 1964 (Table 25). An approximate 45,000 elk days of foraging by 2,500 animals from October 14 through 31 in 1962 declined to 200 to 300 elk arriving, but immediately leaving, by 1966 and 1967. The objective of reducing late migrating Yellowstone groups (major portion of north migratory segment) was probably also achieved by the increased November and proportionately higher hunting removals after 1963. This, and compensating increases in herd segments

Table 25.--Data showing trends in reducing October migrations to refuge winter ranges, reducing late migrating segments that cross Grand Teton Park, and allowing increases in herd segments that migrate through areas outside the park.

Year	Maximum October counts on refuge	Percent removal north migratory segment	Percent of total tracks crossing	
			Grand Teton	Outside
1959	1,550	(Probable	--	--
		(
1960	2,500	(average	58	42
		(
1961	3,500	(over period	--	--
		(
1962	2,500	(13% or less ¹	62	38
1963	728	23	80	20
1964	1,000	19	75	25
1965	375	26	67	33
1966	300	17	66	34
1967	200	17	43	57

¹ Based upon 1964 calculations, season closing dates and 1959 and 1960 male only seasons (Cole, 1965).

summering on and migrating over lands outside park boundaries, may be shown by the track count differences since 1963 (Table 25).

The 1967 track count differences reported by Houston and Yorgason (1968) appeared to result from some migrating groups taking more direct routes to wintering areas instead of converging on Grand Teton Park. This could represent some initial change from behavior adjustments to hunting disturbances, but it may have also resulted from removals of older female elk that perpetuated migration "habits" which previously had high survival value. Some complementary assistance may have also resulted from deferring either-sex seasons, in more accessible hunting areas with low numbers of elk, until the end of the main breeding season (September 21 to October 10). The bugling of adult males associated with harem groups facilitated locating elk groups that would otherwise be difficult to find.

The cooperative program is considered to illustrate that two agencies, with somewhat different responsibilities in public service and philosophies toward land use and wildlife, can effectively work together. A full restoration of elk distributions and migrations to approximate those before 1950 may take at least a decade. A reduction of Grand Teton's role to standby status, with limited hunts when state programs need assistance, may be possible within a few years. Hunting systems on lands outside the park will need to be carefully regulated to retain historical elk distributions and migrations.

ECOLOGY

This section attempts to integrate study findings into an overall account of the relationships between the elk population and its environment. Reconstructions of probable past relationships for comparison purposes were aided by reference literature on ecological principles (Elton, 1927; Allee et al., 1949; and Daubenmire, 1968), Errington's (1946) reviews on vertebrate predation, and by the author's preliminary findings from a study (since June 1967) of a naturally regulated elk population within the west central portion of Yellowstone Park. Other cited literature also aided interpretations on population ecology or predation. Subsections on environmental influences, behavior, and habitat relationships led to final considerations of elk population regulation.

Environmental Influences

Climate and Weather

Climate had long-term influences through its role in directing relatively slow plant succession processes which changed habitat and food conditions. As periodic severe winter weather, it had pronounced short-term effects on the elk population. Deep snow and/or particular crust conditions limited the quantity and/or quality of winter food available to the animals by restricting their movements and foraging actions. These conditions, in combination with low temperatures,

caused energy stresses which influenced population mortality and reproductive rates.

Winter Food and Plant Succession

Winter food represented the most limited source of energy for the studied elk population. Other environmental influences and the distribution and size of the elk population determined the extent to which winter food was utilized or in limited supply.

Elk that freeranged on ecologically complete winter habitats that were complexes of bottomland, swale, and slope areas (or ecological equivalents) seemed unable to progressively reduce snow-covered food sources by their foraging activities alone. The animals maintained what could be considered natural biotic disclimaxes or zootic climaxes (Daubenmire, 1968) on limited snow-free ridgetops and upper slopes. Elk that were concentrated by artificial food sources or were otherwise restricted from using ecologically essential units of winter habitat maintained disclimax conditions that probably would not have occurred with freeranging animals.

Plant succession since the retreat of the last glaciers has undoubtedly increased winter food sources for elk. This process still continues on wintering areas where pioneer and some disclimax stands (from past summer livestock grazing) of big sagebrush are being slowly replaced by more preferred grasses and shrubs.

The replacement of seral willow and aspen stands on elk wintering areas with deep snows has and will continue to represent some reduction in available food. This was partially offset by the reestablishment of successional young stands, but long range trends seemed to be toward progressively less willow and aspen. The replacement of willow stands in areas with lesser snow depths resulted in quantitative gains in herbaceous forage. These gains were partially offset by herbaceous vegetation being less available than willow during most winters. The replacement of aspen by conifers or herbaceous vegetation on areas with lesser snow depths has proceeded to the point where existing remnant stands do not represent a quantitatively significant winter food source for the majority of the elk population. Aspens' role during earlier times with geologically young substrates, an absence of competing vegetation, more frequent fires, and possibly a different climate may have been more significant. Elk and other plant-eating animals hastened the replacement of seral willow and aspen after succession reached advanced stages or stands were reduced to remnant status.

Predators and Scavengers

These roles are not separated because many native meat eaters were both predators and scavengers. Also, relationships appear to exist where scavenging required more efficient predators to make additional kills or contend with food limitations set by scavengers.

Original predator and scavenger populations may have periodically dampened and extended the interval between elk population peaks (see Population Regulation). Reductions in the numbers and/or kinds of predator and scavenger fauna have undoubtedly changed these relationships.

Organized predator control programs during the early 1900's contributed to the near extermination of the mountain lion (*Felis concolor*) and grey wolf (*Canis lupus*) from the Grand Teton and Yellowstone regions. Mountain lions have been sighted in both parks in recent years, but seem unable to increase their numbers. A known 134 wolves were killed within Yellowstone Park between 1916 and 1926. The animals were believed to have been eliminated. However, park records show fairly consistent sightings of single wolves and groups of up to four animals have been reported within Yellowstone Park since 1932. Present low numbers and distributions, almost entirely within Yellowstone Park, precluded the wolf from having a significant influence on the studied elk population. The black bear (*Ursus americanus*) and grizzly bear (*Ursus horribilis*) were also precluded as significant predators by their low numbers or distributions.

The coyote was protected within Yellowstone Park after 1935. Unrestricted shooting or control programs continue to the present on the lands outside national park and elk refuge boundaries. The coyote seemed to have sufficiently secure habitats to remain abundant and in some dynamic relationships with the studied elk population.

Accounts of predators killing elk and other wild animals from Murie (1940), other unpublished records from Yellowstone Park files, and the author's observations of attempted or circumstantially successful (examinations of carcasses, tracks in snow, etc.) predation between 1962 and 1968 suggested that past and present predation on elk did not depart from general principles summarized by Errington (1946). These were that predation was largely limited to some portion of the annual production of young, to animals predisposed to being preyed upon by accidents, sickness and old age, or to animals attempting to inhabit low security level habitats where environmental conditions made them vulnerable.

Elk carrion may have represented a seasonally important food source for golden (*Aquila chrysoetas*) and bald eagles (*Haliaeetus leucocephalus*), ravens (*Corvus corax*) and magpies (*Pica pica*). It, in addition to predisposed adult elk, appeared to represent a food source which could influence coyote and grizzly bear numbers. Murie's (1940) study in Yellowstone Park suggested that direct relationships exist between winter and spring carrion sources and coyote numbers. Studies by Jonkel (1967) suggest that, following summer or fall seasons with poor berry or whitebark pine nut crops, the survival of subadult bears could depend upon their finding carrion or having predisposed animals to prey on after they emerged from hibernation.

Coyote predation on adult elk seemed limited to weak animals unable to stay within groups or very near death. Other elk actively kept coyotes at the margins of larger groups and away from predisposed individuals. Grizzly bears appeared to be more efficient predators through their apparent ability to course and pull down adult elk. Their predation, however, seemed to be largely restricted to newborn young and predisposed or vulnerable animals.

Most early accounts of wolf predation in the Grand Teton and Yellowstone region reflect the predator control attitudes which prevailed into the mid-1930's. Park records since this period (to 1969) are limited to eight reports of one to four wolves feeding on elk within Yellowstone Park and one sighting of a single wolf on the offal from a hunter-killed elk about 3 miles north of Grand Teton. Reports by Olson (1938), Murie (1944), Cowan (1947), Stenlund (1955), Burkholder (1959), Mech (1966), and Pimplott (1967) suggest that wolves could be a more efficient predator on large ungulates than the grizzly. This, however, might only be to the extent that the animals occurred in packs of some optimum size.

Parasites and Disease

The original ecological role of indigenous parasites and diseases probably represented a form of selective "predation" on weak or aged elk and assistance (predisposing) to predation that hastened the deaths

of such animals. Such relationships appeared to still partially exist, but at a relatively inefficient level. Without selective culling by larger predators, predisposed elk (by disease and other agents) commonly persisted over winter or into early spring before dying. Infestations of the winter tick (Dermacentor albipictus), and scabies from a psoroptic mite (Psoroptes sp.) were the most apparent manifestations of disease in the studied elk population.

Other Animals

Present day relationships between elk, other wild animals, and their food sources are illustrated on a portion of a food web diagram (Figure 15). The vertical lines between herbivores and representative forage plants within different habitats illustrate food sources which could, in combination with other environmental influences, limit population size. Dashed sloping lines show dual or multiple use on food sources between different herbivores. Solid lines from herbivores to predators, scavengers, and parasites may represent either specific or general food links.

Significant intensities of dual or multiple use appeared to be limited to areas where the different habitat types or main food sources for wintering elk, moose, or mule deer adjoined each other. Suggested relationships were that differences in food habits and/or environmental conditions influenced distributions and prevented biologically

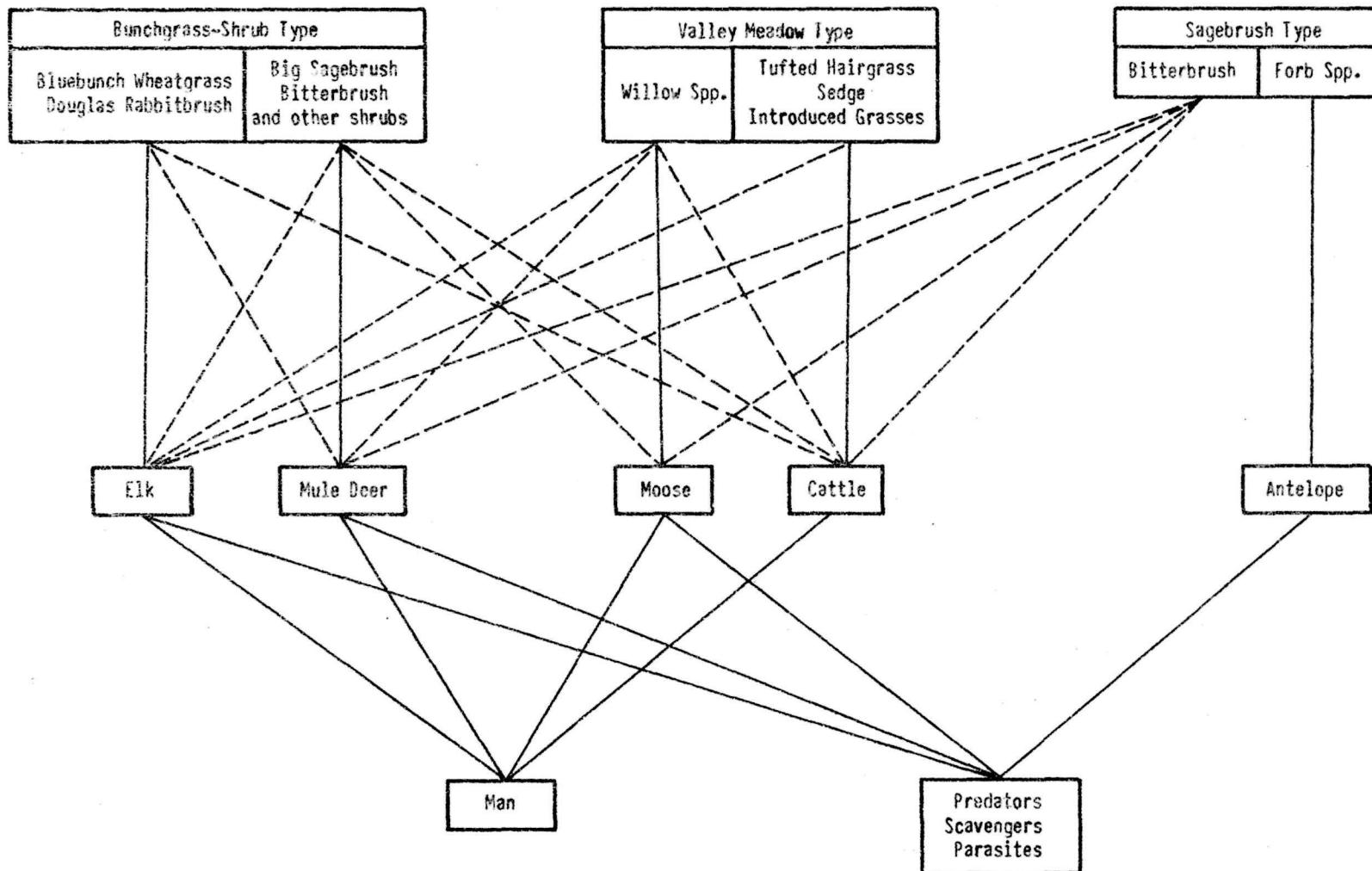


Fig. 15. Food relationships between animals and vegetation types on valley winter ranges.

significant competition where one species could progressively displace the other. Varying degrees of "exclusion" competition occurred on portions of areas or habitats with the greatest overlapping use by different species. Such interspecific competition seemed an essential relationship which prevented one species from appropriating the habitat or food niche of another.

Deep snow (24 or more inches) appeared to allow moose to "outcompete" and ultimately preclude elk from using significant portions of their main willow food sources. Lesser snow depths permitted elk to utilize willow in conjunction with their more abundant herbaceous food sources on or adjacent to bottomlands and ultimately "outcompete" moose. Wintering mule deer contributed to overlapping use, but appeared to persist because of abundant shrub food sources on slope areas and their ability to utilize big sagebrush as a main food item. Snow conditions variably influenced the availability of shrub food sources on slopes and the extent of overlapping use by different herbivores.

Small numbers of pronghorn antelope that were a remnant of a larger population that summered within Grand Teton Park at the turn of the century. Some animals occasionally attempted to winter within the park. Deep snow made them vulnerable to the substantial coyote population which appeared to be largely sustained by elk carrion.

Small bands of bighorn sheep (Ovis canadensis) stayed within the Teton Mountain range yearlong. Another group wintered on the east side of the National Elk Refuge. These animals mainly wintered on and adjacent to rock outcrop areas. Such habitat areas were undoubtedly more extensive when the region was geologically younger and/or the climate was sufficiently warm to permit the winter use of sites which are now within deep snow zones and/or tree covered.

A warmer and drier climate occurred between 9,000 and 4,500 years ago (Kind, 1965). It seems probable that climatic change and ecological succession, involving elk as well as other faunal species, since this period have contributed to the present relict status of the bighorn. Archeological excavations bordering Yellowstone Park show a race of mountain dwelling Indians may have subsisted mainly on bighorn sheep at least 4,500 years ago (Wedel et al., 1968). Modern man's influences may have additionally contributed to the bighorn's present status by eliminating more vulnerable groups from accessible habitat areas and grazing domestic stock on their winter ranges.

Elk relationships to predators, scavengers, parasites and diseases have been discussed. In addition to their use of the elk as a host, certain insects strongly influenced the distributions and movements of elk and thereby their use of food sources (see Elk Habits and Habitat Use).

Man

It seems unlikely that the summer visits of Indians, early trappers, or even the first few settlers in the Jackson Hole region up to about 1910 had significant impacts on the large elk population or changed environmental conditions for the animals. This was not the case after human settlement reached levels where the elk's historical wintering areas were almost completely appropriated for domestic stock grazing or hay raising. Figure 16 illustrates the extent to which modern man has become part of the elk's environment. Human developments, agriculture and hunting, in combination with the practice of winter feeding the animals, have become major influences on present day elk populations.

Developments and Winter Ranges

The use of valley lands south and north of the town of Jackson for hayfields, livestock pastures, ranch or home sites, and more recently, a golf course has led to the present elk herds being restricted to wintering on about 60 percent of the land described as their winter range in 1911 (Figure 6). No lands were specifically set aside for wintering elk until 1913 when 2,800 acres were designated as the National Elk Refuge. About 1,760 acres were privately donated in 1927. Purchases and administrative actions between 1935 and 1950 led to the refuge reaching its present 22,700 acre size.

Refuge, adjoining national forest, and enclosed state lands presently make up a 37,500 acre unit (Houston, 1968a). This, a 18,700

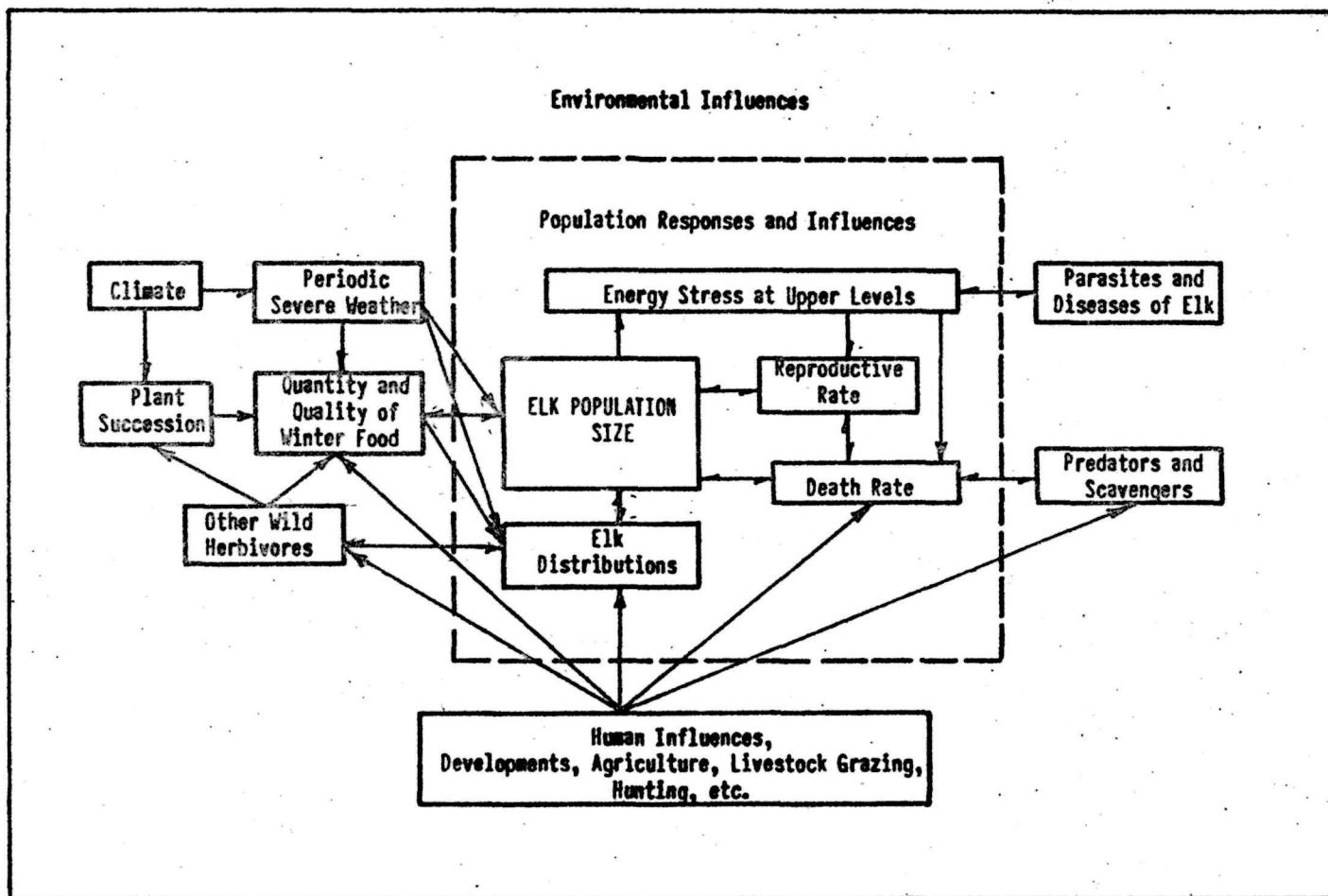


Fig. 16. Elk population regulation.

acre unit within Grand Teton south of Ditch Creek, and an adjoining 4,000 acre slope area to Horsetail Creek total about 60,000 acres and make up the largest single block of publicly owned historical winter range available to the elk in the Jackson Hole valley. Additional wintering areas within Grand Teton Park north of Ditch Creek and the adjoining Buffalo River valley total over 7,000 acres.

Preble (1911) described the extensive "marsh" bottomlands along Flat Creek above the town of Jackson as a "favorite haunt" for wintering elk. He reports other bottomland or slope areas as wintering 2,000 or 3,000 elk, harboring good sized herds, a few hundred, or a few animals. Greater numbers of elk appeared to winter where the animals had extensive bottomlands in addition to upland slopes. The Flat Creek bottomlands north of the town of Jackson occur within present refuge boundaries. Other comparable bottomland areas that Preble mapped as elk winter range lie west of present refuge boundaries and south of the town of Jackson. These have become almost entirely privately owned and wintering elk are either restricted to using small State owned feed grounds or slope areas bordering the main valley. Less extensive bottomland areas, formally mapped as winter range, occur in Grand Teton's Spread Creek-Uhl Hill area, the adjoining Buffalo River valley, within the Gros Ventre River flood plain along the north boundary of the refuge, and scattered along the Snake River in and outside Grand Teton Park. These bottomlands were or still are variously used for hay raising and/or livestock grazing.

Livestock Grazing

Approximately 11,000 a.u.m.'s (one animal unit grazing one month) of grazing by over 4,000 cattle and horses over a May 1 to November 15 period were acquired when national monument lands were added to Grand Teton Park in 1950. Progressive expirations of lifetime leases and transfers of land could result in the elimination of large scale livestock grazing within the park some time after the year 2,000. Extensive open range grazing in the park has been progressively transferred to fenced pastures on the east side of the Snake River since 1958. Some pastures included important wintering areas for elk, mule, deer, and moose.

Figure 15 shows cattle had the capacity to compete with elk, mule deer, or moose. This appeared to reach significant levels in the park's Spread Creek-Uhl Hill areas. Here, 1963-1966 measurements showed late summer and fall cattle grazing reduced main winter food sources for elk and mule deer (see Effects on Habitats). Elk shifts to use adjoining moose habitat and food sources were pronounced in 1964. Rancher cooperation was partially secured after 1964 to move cattle to irrigated bottomland pastures before they started to use elk and mule deer winter food sources. Cattle and horse grazing on pastures within the southern portion of Grand Teton Park would only cause conflicts if large numbers of elk again used the historical winter ranges between Ditch Creek and the north boundary of the refuge.

Hunting

Man's hunting directly influenced elk population death rates and distributions (Figure 16). Indirect influences resulted from hunting other wildlife that might occur in some relationship with the elk as a competitor, scavenger, or predator. Elk distributions were probably influenced by removing particular animals from herd groups, conditioning survivors, and by simple avoidance responses to hunting disturbances which caused animals to move to superior escape habitats or areas where they were undisturbed. Martinka (1969) reported on the 1964 movements of marked elk on and across Grand Teton and the National Elk Refuge and concluded hunting encouraged movements to and discouraged movements from areas closed to hunting. The results from the purposeful use of hunting to halt early habitual movements to the refuge and its unintended effects on the historical distributions and migrations of other elk groups are discussed in the Fall Migration section.

Artificial Feeding

Differences between high pregnancy rates and calves at heel, the high spring mortality of previous year calves following severe winters, and the comparative appearance of animals on and off refuge feedgrounds suggested that artificially fed hay was either a nutritionally inadequate diet for subadults and pregnant females or the

feeding process itself increased overwinter energy stresses to levels which reduced the net quality of the diet.

The greater exposure of animals on open feed grounds to weather influences may have been partially involved. Moen (1968) showed wind in combination with freezing temperatures could progressively reduce the effectiveness of both high quality and maintenance diets in balancing energy losses in white-tailed deer. Smaller animals were most affected. This basic surface to mass relationship suggests that the reduced opportunity for calf elk on open feed grounds to obtain shelter from wind could partially account for their usual less thrifty appearance (by late winter) than calves in free-ranging groups.

The establishment of energy-conserving social relationships (peck orders) between different aged animals or dominant individuals appeared impossible when large numbers of elk were concentrated on feeding grounds. The subtle and overt agonistic behavior that occurred with the daily crowding or intraspecific competition of 1,000 to 4,000 elk on feed lines could be expected to expend energy that reduced the effectiveness of a diet. The consequences of crowding were most apparent in the rapid breakdown of female-calf associations. Large groups of calves that were obviously not at heel concentrated together on feed grounds. Calves in smaller free-ranging groups off feed grounds appeared to be associated with maternal females through the winter.

Refuge records (Table 26) indicate that approximately equal winter herds have been maintained from 1912, when large groups of elk (7,250 animals) were first fed, to recent years. Periodic low numbers up to about 1945 probably resulted from initially "high" mortality occurring both on, and after elk left, the refuge as well as the presence of herd segments that did not use feed grounds consistently. Undetected mortality on spring ranges apparently substituted for mortality on feed grounds after 1945. Elk also used feed grounds more consistently. After 1945, higher hunting removals, periodic "high" mortality off feed grounds and compensating reproductive increases apparently maintained winter herd numbers between about 6,000 and 11,000 animals (7,600 average).

Table 26 also shows artificial feeding for 3 to 4 months on relatively small areas has had the end result of partially substituting hay diets for natural foods on extensive areas previously used by the animals. This was undoubtedly necessary because of the limited wintering areas set aside for the animals before 1935, and a subsequent state damage law that tended to make artificial feeding less costly and controversial than continuous investigations and payments for elk damage.

The particular conditions that may have contributed to the refuge winter herd maintaining its numbers for over 50 years are:

Table 26.--Numbers of elk fed hay diets on the refuge and recorded losses, 1912-1967.¹

Periods	Number of elk fed			Ave. tons hay fed (thousands)	Recorded losses		
	Maximum	Minimum	Average		Average	Maximum	Pct.
1912-15	7,300	4,000	5,600	0.5*	--	--	
1916-20	10,000	3,000	7,000	1.0	311	15	
1921-25	5,500	3,400	4,400	0.8	210	21	
1926-30	7,500	6,000	6,300	1.5*	103	5	
1931-35	9,500	3,000	7,100	1.8**	180	7	
1936-40	9,500	4,000	6,300	0.9*	98	3	
1941-45	10,000	5,200	8,000	1.4*	366	11	
1946-50	8,700	6,500	7,400	1.2*	120	2	
1951-55	9,500	7,200	8,700	1.5	103	3	
1956-60	11,000	6,000	7,400	2.2	71	1	
1961-65	8,200	5,800	7,200	3.0	81	2	
1966-67	7,400	6,600	7,000	2.5	32	trace	

¹ From National Elk Refuge records.

* No feeding one year within period.

** No feeding two years within period.

1. Initial high mortality of calves could have led to some selection for a population more adapted to feed ground conditions.
2. Abundant natural food sources became available to the animals as lands were added to the refuge after 1935.
3. Artificial feeding for 3 to 4 months provided a relatively consistent subsistence diet that could be supplemented by abundant natural foods adjacent to feed grounds.
4. Feed grounds were located on an extensive 6.5-mile long by 3-mile wide fertile bottomland flat where trampling and close cropping of vegetation was accommodated without causing destructive overuse of upland sites away from their immediate vicinity.

Man's continued routine feeding of the refuge elk herd after extensive historical wintering areas became available for their use appears to be partly tradition and partly the very real prospect of damage claims from the relatively few ranches scattered within or bordering historical wintering areas inside Grand Teton Park. A possible solution to the damage claim problem on park lands could involve securing scenic easements until present leases expire.

Behavior Relationships

Behavior that appeared to be caused by external environmental influences or relationships within the elk population itself is listed

in Table 27. Spring dispersals off feed grounds, November migrations that occurred with particular snow conditions, and aggregations of animals into large groups on refuge feed grounds appeared to be food-linked responses.

Movements off the refuge to Grand Teton spring ranges and into Yellowstone Park seemed initially to be a response to reduced environmental resistance; by mid-May, some overriding attempts by females to reach calving areas. These, the dispersals off feed grounds and November migrations, tend to show behavior that resulted from a progressive relaxing and return of severe weather conditions that influenced the availability of food or restricted movements along a general elevational gradient (the Snake River drainage). It appeared that elk would remain in scattered distributions on summer home ranges in the absence of weather conditions that reduced the availability of their food. Weather severity along elevational gradients resulted in the animals moving to and concentrating on lower elevation areas where essential members of the population survived the severest winters. Moderating weather allowed reverse movements and a return to scattered distributions on summer home ranges.

Delayed movements of a portion of the elk population onto high elevation ranges, August dispersals into forest types, and the formation of harem groups represented behavior linked to reproduction.

Table 27.--Summary of elk behavior caused by external environmental influences and intraspecies relationships.

Behavior	Indicated causes
Spring dispersals off refuge feedgrounds	Moderating weather exposing snow-free areas and initiating new grass growth.
Movements off refuge to Grand Teton spring ranges	Moderating weather reducing snow depths which restrict movement or, by mid-May, attempts by females to reach calving areas.
Variable use of calving areas and rates of movement through mountain areas	Variable late May-early June snow conditions in mountain passes.
Delayed movements of female/calf groups onto high elevations	Initial care-dependency relationships between female and young.
Mid-summer aggregations on high elevation ranges	Molesting insects.
August dispersals from high ranges into forest types	Female avoidance to early sex-linked behavior of adult males.
Mid- to late October movements of some elk to refuge wintering areas	High levels of human disturbances on Grand Teton fall ranges that have limited escape cover for large elk groups.
November aggregations and migrations from mountain areas	Snow conditions restrict access to food along an elevational gradient.
Rapid and direct movements over valley early winter range areas to the south half of the refuge	Human disturbances and hunting in areas with limited forest cover.
Aggregations of large groups of elk on small feedground areas	Conditioned habitual use of easily obtained hay diets by adult animals and leader-follower relationships.

Newborn calves delayed the movements of maternal females and other associated animals to the most distant Yellowstone summer ranges and to high elevations until July. Subtle attempts by previously segregated adult male elk to associate with groups of females and calves (observed as early as August 9) contributed to breaking up summer group associations and caused dispersals into forest types. Overt displays by adult males and attempts to collect females in harems (rutting behavior) were observed as early as August 13 and commonly after this date. Field records showed all male elk older than yearlings had started or completed removing the velvet from their antlers by August 25. The process was observed to start as early as August 13. Attempts to hold females in harem groups appeared to be most successful between mid-September and mid-October. Over the mid-August through October period, the size of elk groups in mountain areas averaged about six animals ($N = 4999$). Yearling males were usually excluded from harem associations of females and calves that were attended by an adult male.

Classifications obtained by Martinka (1965) showed late August through October group sizes in valley areas were much larger, averaging about 30 animals ($N = 12,936$). An average group size of 11 animals was observed during the mid-September to mid-October peak of breeding activity. These higher group sizes probably resulted from fewer adult males being initially present in valley areas and a greater tendency for harem groups to aggregate on extensive fall range areas with limited forest cover.

October migrations of Grand Teton elk to refuge winter ranges appeared to represent behavior responses to viewing by park visitors and some illegal hunting disturbances on large elk groups (200 to 500 animals) within low security level habitats (outwash plain with limited forest cover). After 1964, old roads within the western portions of Grand Teton were closed to provide blocks of fall range where large elk groups could be seen from vista points, but not disturbed by too close an approach. These roads had either penetrated major fall range areas for large elk groups or allowed untolerated approaches between foraging animals and their forest cover. The road closures, in combination with limited permit hunting on eastern portions of Grand Teton and the National Elk Refuge, greatly reduced early October migrations from the western portions of the park.

As discussed in the Habitat Use section, aggregations of large numbers of elk at high elevations in mountain areas were caused by molesting insects. The equivalent behavior response for elk in Grand Teton valley areas or in groups scattered at low elevations within mountain areas was to retire into forest types or bed in dense herbaceous vegetation in meadows. This would suggest that high elevations were not a deciding or critical factor in the animals' establishment of summer home ranges. Brazda (1953) sampled molesting insect densities which indicated elk would obtain greater relief at high than low elevations. An additional relationship may have been that

large aggregations of elk afforded relief. Allee et al. (1959) cites an account that aggregations of at least 300 to 400 reindeer (Rangifer sp.) permitted herds to remain intact under warble fly attacks.

The suggested relationship was that elk largely influenced their own scattered distributions when environmental stresses were minimal. These were variably maintained by matriarchal care-dependency and leader-follower associations with a dominant female elk; by mutual avoidance, subtle and overt agonistic behavior that maintained dominance-subordination relationships within and between associations; and by sexual relationships between adult males and females. The latter involved either mutual or female avoidance behavior during periods other than the breeding season. Terminology follows Etkin (1964).

Aggregations of elk from scattered summer distributions and fall migrations occurred in relation to overriding environmental influences and coincident subjugations of some dominant females into leader-follower relationships. Subjugations may have resulted from some lessening of female dominance apart from home range "territories" used for the care of young. Aggregations of female and subadult elk on wintering areas apart from feedgrounds probably re-formed as variably sized matriarchal associations with both care-dependency and leader-follower relationships. These re-established social order with energy conserving dominance-subordination (peck order) relations.

Off feedgrounds, adult males usually wintered in loosely organized herds socially apart from matriarchal associations. Such segregations of adult male elk apart from groups of females, calves, and yearling males were apparently the rule in early day elk populations (Preble, 1911).

Large groups of female, subadult, and adult male elk on feedgrounds appeared to represent aggregations where social relationships progressively deteriorated. This may have resulted from daily occurrences of agonistic behavior between large numbers of variably dominant and subordinate elk on feed lines. The establishment of energy conserving peck orders was precluded and led to an early dissolving of maternal care-dependency relations (see Artificial Feeding). Moderating environmental conditions in spring, in combination with the re-establishment of maternal-care relationships, led to the re-establishment of scattered distributions on summer home ranges.

Aggregations of social groups did occur from foraging encounters on other than wintering areas. These were usually temporary or occurred under conditions where close crowding to obtain food was not necessary. Aggregations from escape encounters occurred in response to varying intensities of human disturbances and molesting insects. Male sexual behavior appeared to cause dispersals from aggregations and either temporary or lasting disruptions of summer matriarchal associations.

Habitat Relationships

Elk have apparently persisted for thousands of years in the Grand Teton and Yellowstone regions over a wide range of environmental changes which are still occurring. Vegetation changes have been short term and cyclic from fire, biotic influences, and variable growing conditions; or directional from developing soils, stream cutting, and climatic change. Selection pressures for the "most fit" plant and animal species have undoubtedly occurred and will continue. It seems unlikely that elk would have persisted if the animals were able to progressively deplete their main food sources which, in combination with other influences, determined their numbers (i.e., had population consequences).

Winter habitats that were interspersions of different physiographic sites and/or vegetation types provided increased opportunities for an elk population to remain in some dynamic balance with its food sources (homeostasis). These ecologically complete habitats had carrying capacity relationships where "the whole was greater than the sum of its parts." The elk's variable use of different habitat units, general food habit, protection from snow, and the capacity of native plants to withstand periodic heavy use appeared to preclude free-ranging animals from progressively depleting their main winter food sources. The density-influenced mortality of animals with low energy reserves also helped to maintain elk populations in balance with their food sources (see Population Regulations)

As biotic agents, elk influenced the rate at which late stages of seral vegetation were replaced, maintained relatively stable biotic disclimaxes on limited sites where their effects were either without population consequences, or were incidental to the use of food sources that had population consequences. They also occurred in some dynamic relationship with other native herbivores through "exclusion" or interspecific competition that retained a mixed species fauna within different food or habitat niches.

Exceptions to these apparently natural relationships occurred on upland areas adjacent to refuge feedgrounds and wildlife wintering areas additionally grazed by domestic stock. Here, animal concentrations and/or consistent heavy or dual use of vegetation appeared to intensify disclimax conditions or cause seral vegetation to be replaced at a faster than "normal" rate.

These interpretations of free-ranging elk relationships to their winter habitats may only apply to other areas with equally variable and rigorous winter weather. They would have limited to almost no application where human influences restricted or precluded elk from using portions of a winter habitat (e.g., bottomlands, slopes, etc.) which were essential to homeostasis. What may be shown is that interpretations of elk habitat relationships require considerations of natural successional processes, the ecological completeness of winter habitats, and distinctions between food sources which do or

do not have population consequences. Natural biotic effects or sucessional changes would not require corrective management within a national park.

Population Regulation

The logistic curve relationship between population growth and environmental resistance may have been first expressed by Verhulst in 1883 (Allee, et al. 1949). Accumulated knowledge since this date further establishes that animal populations occur in some equilibrium (mean numerical stability) in the absence of environmental changes that consistently cause more or less resistance to population growth. Environmental changes which consistently offer less environmental resistance permit upward trends in population numbers. Consistently more environmental resistance results in downward population trends.

A regulating influence was considered to directly or indirectly cause deaths, or change population reproduction, or survival rates. The complementing influences from some animals emigrating from or immigrating into a population is recognized. The probable regulatory process for past as well as present populations is presented for comparison purposes.

Past Populations

Accumulated knowledge on the organization of life in natural communities tends to assure that past elk populations were regulated to the extent that they could not, by themselves, progressively

deplete food sources which limited their numbers. This study suggests that the animals could have temporarily reduced the amount or quality of their own food sources as part of a natural regulatory process, reduced or maintained some food sources that did not limit their numbers as natural biotic disclimaxes, and accelerated late stages of plant succession to either increase or decrease their total food supply.

Intraspecific competition for available food and environmental influences from winter weather, predators, scavengers, and diseases probably interacted to lower the numbers in early-day elk populations in the following manner: When populations were at upper levels in relation to their available winter food, intraspecific competition intensified energy stresses. These stresses directly or indirectly caused the deaths of elk with the lowest energy reserves and sometimes lowered the subsequent year's reproductive success. The deaths of diseased and other energy-stressed animals were hastened by the combined effects of predators and scavengers.

Severe winter weather per se periodically caused higher than usual deaths, or what could be considered additional density-independent mortality, by increasing intraspecific competition, energy stress, and the efficiency of predators. These additional deaths were also animals with low energy reserves. Subsequent winters with less severe weather or intraspecific competition

permitted elk populations to compensate for the deaths of animals with low energy reserves and return to higher numbers. Compensations resulted from increased reproductive success and survival or the process Errington (1946) calls "compensatory trends."

The reports of "high" winter mortality in the Jackson Hole herd at 4 to 6-year intervals during 1882, 1887, 1891, 1897, and 1911 (Preble, 1911; Anonymous, 1915; Sheldon, 1927; Brown, 1947) suggest that early day predator populations did not prevent the elk from contending with the regulatory influences from intraspecific competition for food or periodic severe winters. This should not be interpreted that predation on elk populations was without ecological significance. Original predator populations may have reduced the intensity of intraspecific competition within an elk population during more severe winters. The extent to which this occurred would have extended the interval between and dampened elk population fluctuations. The compensatory trend process (Errington, 1946) could be expected to compensate for periodically higher than usual mortality from predation.

Present Populations

Intraspecific competition for food and environmental influences from man, winter weather, a limited predator-scavenger fauna, and disease acted to lower numbers in present elk populations (Figure 16).

The regulatory process mainly differed from that on past populations to the extent that man increased or decreased the intensity of natural regulatory influences. This resulted from his artificially feeding the animals, displacing the original predator-scavenger fauna, and changing herd distributions so as to reduce total food sources.

When elk on winter feedgrounds were at upper levels (apparently a wide range) in relation to the available energy from their artificial diets and adjacent food sources or the effects of periodic severe weather, intraspecific competition increased energy stresses. These stresses directly or indirectly caused the deaths of subadults and older adults with the lowest energy reserves and sometimes lowered the subsequent year's reproduction. The deaths of diseased or energy-stressed animals were not significantly hastened by the remnant predator fauna which was mainly restricted to preying on newborn elk. Population increases back to higher levels, by compensating reproduction and/or survival, were influenced by hunting removals and other human influences.

The density-influenced or periodically higher density-independent deaths of animals with low energy reserves would not represent a loss of biologically essential population members and would be predestined to occur to the extent that elk populations were self-regulated (intraspecific competition) in relation to their available

winter food. Such mortality would not occur to the extent that hunting removals could substitute for density-influenced deaths.

From 1962 through 1967 about 700 to 800 elk were estimated to consistently winter without using artificial food sources. These animals occurred in scattered distributions on historical winter areas on the refuge, Grand Teton Park, and adjoining national forest lands. Their numbers were variously regulated by interspecific competition, weather influences, hunting, and competition with domestic livestock. Small elk groups, that remained within Grand Teton areas closed to hunting or arrived on park winter ranges after hunting seasons were closed, were largely self-regulated by intra-specific competition for food and weather influences.

Generally high hunting kills from 1940 through the late 1950's coincided with progressive declines in the numbers of elk that freeranged off refuge feedgrounds. This reduction in animals, which were partly or wholly on different food sources, may preclude presently distributed winter herds from reaching 1955-56 and earlier year population peaks of 10,000 to 11,000 animals. Sustained hunting removals that approximate herd increase rates and yearly artificial feeding may additionally restrict winter herd numbers from fluctuating outside the general 6,000 to 8,000 range that has prevailed over severe, average, and mild winters since 1961. Herds could be expected to occasionally fluctuate to higher levels with lower or less consistent

hunting removals. The extent to which artificial feeding prevented subadults and/or other population members from freeranging in scattered distributions and obtaining more adequate diets could also restrict fluctuations to higher levels.

Man's actions in restricting elk from freeranging were not always unintentional when the animals' historical winter range was largely privately owned and used for livestock grazing or hay raising. Transfers of land, purchases, and administrative withdrawals have progressed to where the refuge herd could be allowed to freerange over a 60,000-acre block of this historical winter range. This could conceivably replace all artificial feeding.

Consistent artificial feeding of the refuge winter herd may result in a more apparent than real lowering of overwinter mortality rates by maintaining low proportions of vulnerable subadults in the population and only deferring until spring what Preble (1911) and others have considered "high" mortality from severe winters (about 15 to 20 percent of herd numbers, or a large portion of the calves). Comparable mortality of subadults and other elk with limited energy reserves appears to have occurred periodically over a wide range of population size up to the most recent severe winter of 1961-62. Such partially density-independent mortality would preclude maintaining highly stable elk populations.

If the present refuge herd was largely self-regulated as a result of low hunting removals and was not artificially fed, its winter numbers could conceivably fluctuate within a 5,000 to 9,000 range. If artificial feeding supplied more energy to subadults and pregnant females than they could obtain by freeranging (this needs to be demonstrated), the population might fluctuate within a 6,000 to 9,000 range; if it did not, a 5,000 to 9,000 range. The latter approximates the usual range of winter herd numbers that occurred before the higher hunting removals of the 1950's (Table 23 and Figure 8).

With hunting removals that attempted to reduce intraspecific competition by approximating average population increase rates, the winter numbers of a free-ranging herd might fluctuate between 5,000 and 8,000 animals. This compares with 6,000 to 8,000 fluctuations between 1961 and 1967 which may or may not have been maintained at higher levels by artificial feeding. These figures should be considered approximations which are mainly used to illustrate suggested relationships. Fluctuation ranges could vary with unusual sequences of winter weather, variable hunting removals, or changes in artificial feeding practices.

Man's hunting since 1955 appears to have become somewhat more efficient than original predator-scavenger complexes in preventing extreme fluctuations in elk numbers. It has not prevented the elk

population from being additionally regulated by periodic severe weather and intraspecific competition. Complete substitutions of hunting for all natural mortality do not appear possible because severe weather influences on the availability of food and on subadult elk or other animals with low energy reserves were not completely density-dependent within the full range of population numbers accommodated by variation in the winter environment.

Man was obviously less successful than the original predator fauna in allowing the elk population to maintain its numbers and distributions in relation to suitable habitats and food sources. This resulted from his more efficient and less restricted (to predisposed and vulnerable elk) hunting reducing elk population groups that used particular habitat areas and forage sources. Conditioned avoidance behavior appeared to additionally restrict elk from using extensive wintering areas with abundant forage sources.

DISCUSSION

The National Park Service, Wyoming Game and Fish Commission, U. S. Forest Service, and Refuge Branch of the Bureau of Sport Fisheries and Wildlife are all governmental units responsible to the public and assigned primary purposes by law. The situation where all are concerned with the same elk requires mutual respect for each agency's assigned responsibilities and cooperative actions for the proper protection and management of the elk.

Purposes of Parks

A 1942 treaty with 17 other countries established Grand Teton and Yellowstone National Parks as part of an international system for "nature protection and wildlife preservation in the Western Hemisphere." Interpretations within the framework of this treaty and the park's enabling legislation led to their being designated "natural areas." As such, they are to preserve natural environments and their biota, and provide opportunities for visitors to view and appreciate scenery and native plant and animal life as it would have occurred in primitive America. This amounts to the preservation of natural ecosystems for their scenic, educational, cultural, and scientific values.

Special management and protection measures are applied to provide opportunities for park visitors to view, photograph, or obtain an appreciation for natural scenery and wildlife. Roads, vista turnouts,

scenic loops and trails are designed to provide access to scenic features and locations where wildlife can be seen. Interpretive signs, printed matter, museum displays, scheduled talks, and tours assist visitors to inform themselves or be informed to any extent they desire. Having park areas closed to hunting increases opportunities to see and photograph wild animals such as the elk and retains roadside animals most seen by visitors.

Purposes of Other Agencies

The Wyoming Game and Fish Commission, U. S. Forest Service, and Bureau of Sport Fisheries and Wildlife, Refuge Branch, have objectives that are primarily directed toward either managing the elk or their habitat so as to provide recreational hunting. The additional value of providing visitors to the Jackson Hole area with the opportunity to see and photograph the animals is recognized, along with realization that this is best provided for under park conditions or on the refuge during the winter season.

The controlled hunting of migratory elk on lands outside Grand Teton and southern Yellowstone boundaries poses no serious conflict with the primary purposes of the parks. Severe winters, distances from human population centers, large blocks of roadless wilderness, and the variableness of fall migrations tend to make close management of these elk for high sustained hunter harvests difficult. With the present large elk population, the objective of having hunting substitute for all other mortality appears unobtainable. Short of taking

and holding the population to substantially lower levels, numerically high mortality of subadults, adult males, and old animals will occur during and after more severe winters. Such mortality appears to be rapidly compensated for by increased reproduction and survival in subsequent years. An important requirement in management may be that hunting is regulated so that it does not progressively reduce more vulnerable population segments.

The Future

The dual use of wild animals for their scenic and other values on park lands and for recreational hunting when they move outside boundaries will continue to require close cooperation between the National Park Service and other agencies for the foreseeable future. The use of portions of Grand Teton National Park to assist in the overall elk management program, when necessary, may require additional refinements to reduce conflicts with increasing numbers of fall and winter visitors. This could involve qualifying applicants for park permits for marksmanship, animal identification, and a knowledge of regulations for the specific purposes of obtaining needed control with fewer hunters and illegal kills of other wildlife.

The cooperative management program to restore historical elk migrations and distributions is expected to progressively reduce the need for large scale hunting programs within Grand Teton and allow desired elk kills to be obtained by hunting outside park boundaries.

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Appendix I

Common and scientific names of plants.

Common name	Scientific name
Agoseris	<u>Agoseris</u> spp.
Alpine timothy	<u>Phleum</u> <u>alpinum</u>
Arnica	<u>Arnica</u> spp.
Aspen	<u>Populus</u> <u>tremuloides</u>
Aster	<u>Aster</u> spp.
Astragalus	<u>Astragalus</u> spp.
Balsam poplar	<u>Populus</u> <u>balsamifera</u>
Balsamroot	<u>Balsamorrhiza</u> <u>sagittata</u>
Bastard toadflax	<u>Comandra</u> <u>pallida</u>
Bedstraw	<u>Galium</u> <u>boreale</u>
Bentgrass	<u>Agrostis</u> spp.
Bitterbrush	<u>Purshia</u> <u>tridentata</u>
Bluegrass	<u>Poa</u> spp.
Blue spruce	<u>Picea</u> <u>pungens</u>
Brome	<u>Bromus</u> spp.
Brome, mountain	<u>Bromus</u> <u>marginatus</u>
Buckwheat	<u>Eriogonum</u> spp.
Buttercup	<u>Ranunculus</u> spp.
Chickweed	<u>Cerastium</u> <u>arvense</u>
Chokecherry	<u>Prunus</u> <u>virginiana</u>
Cinquefoil	<u>Potentilla</u> spp.
Cinquefoil, shrubby	<u>Potentilla</u> <u>fruticosa</u>
Collomia	<u>Collomia</u> <u>linearis</u>
Coneflower	<u>Rudbeckia</u> <u>occidentalis</u>
Dandelion	<u>Taraxacum</u> spp.
Douglas-fir	<u>Pseudotsuga</u> <u>menziesii</u>
Elk thistle	<u>Cirsium</u> <u>foliosum</u>
Engelmann spruce	<u>Picea</u> <u>engelmanni</u>
Fireweed	<u>Epilobium</u> spp.
Flax	<u>Linum</u> <u>lewisii</u>
Fleabane	<u>Erigeron</u> spp.
Fringed sagewort	<u>Artemisia</u> <u>frigida</u>
Gentian	<u>Gentian</u> spp.
Geranium	<u>Geranium</u> spp.
Giant wild rye	<u>Elymus</u> <u>cinereus</u>
Goldenrod	<u>Solidago</u> spp.
Gooseberry	<u>Ribes</u> spp.
Groundsel	<u>Senecio</u> spp.

Appendix I (continued)

Common and scientific names of plants.

Common name	Scientific name
Hackelia	<u>Hackelia floribunda</u>
Hairbell	<u>Campanula rotundifolia</u>
Hawksbeard	<u>Crepis spp.</u>
Honeysuckle	<u>Lonicera spp.</u>
Horsebrush	<u>Tetradymia canescens</u>
Horsetail	<u>Equisetum spp.</u>
Huckleberry, dwarf	<u>Vaccinium scoparium</u>
Huckleberry, large	<u>Vaccinium membranaceum</u>
Idaho fescue	<u>Festuca idahoensis</u>
Junegrass	<u>Koelaria cristata</u>
Kinnikinnick	<u>Arctostaphylos uva-ursi</u>
Leptodactylon	<u>Leptodactylon pungens</u>
Limber pine	<u>Pinus flexilis</u>
Little sunflower	<u>Helianthella spp.</u>
Lodgepole pine	<u>Pinus contorta</u>
Lousewort	<u>Pedicularis spp.</u>
Lupine	<u>Lupinus spp.</u>
Menziesia	<u>Menziesia ferruginea</u>
Microseris	<u>Microseris spp.</u>
Mountain bluebell	<u>Mertensia ciliata</u>
Narrowleaf cottonwood	<u>Populus angustifolia</u>
Needle-and-thread	<u>Stipa comata</u>
Needlegrass	<u>Stipa spp.</u>
Oatgrass	<u>Danthonia spp.</u>
Oniongrass	<u>Melica spectabilis</u>
Oregon grape	<u>Berberis repens</u>
Pachystima	<u>Pachystima myrsinifolia</u>
Parsley	<u>Lomatium spp.; Perideridia spp.</u>
Phlox	<u>Phlox spp.</u>
Pinegrass	<u>Calamagrostis rubescens</u>
Pipsissewa	<u>Chimaphila umbellata</u>
Plantain	<u>Plantago spp.</u>

Appendix I (continued)

Common and scientific names of plants.

Common name	Scientific name
Rabbitbrush, Douglas	<u>Chrysothamnus viscidiflorus</u>
Rabbitbrush, rubber	<u>Chrysothamnus nauseosus</u>
Red dogwood	<u>Cornus stolonifera</u>
Reedgrass	<u>Calamagrostis</u> spp.
Ricegrass	<u>Oryzopsis hymenoides</u>
Rose	<u>Rosa</u> spp.
Rush	<u>Juncus</u> spp.
Russet buffaloberry	<u>Shepherdia canadensis</u>
Selaginella	<u>Selaginella</u> spp.
Sagebrush, big	<u>Artemisia tridentata</u>
Sagebrush, low	<u>Artemisia arbuscula</u>
Sagebrush, silver	<u>Artemisia cana</u>
Sagebrush, threetip	<u>Artemisia tripartita</u>
Sedge	<u>Carex</u> spp.
Sedge, Geyer's	<u>Carex geyeri</u>
Serviceberry	<u>Amelanchier alnifolia</u>
Silverberry	<u>Eleagnus commutata</u>
Snowberry	<u>Symporicarpos</u> spp.
Spiraea	<u>Spiraea betulifolia</u>
Snowbush	<u>Ceanothus velutinus</u>
Strawberry	<u>Fragaria</u> spp.
Subalpine fir	<u>Abies lasiocarpa</u>
Sweetclover	<u>Melilotus officinalis</u>
Tall larkspur	<u>Delphinium occidentale</u>
Thinleaf alder	<u>Alnus tenuifolia</u>
Tufted hairgrass	<u>Deschampsia caespitosa</u>
Trisetum	<u>Trisetum</u> spp.
Valerian	<u>Valeriana</u> spp.
Violet	<u>Viola</u> spp.
Waterleaf	<u>Hydrophyllum capitatum</u>
Wheatgrass	<u>Agropyron</u> spp.
Wheatgrass, bluebunch	<u>Agropyron spicatum</u>
Wheatgrass, slender	<u>Agropyron trachycaulum</u>
Wild barley	<u>Hordeum</u> spp.
Winterfat	<u>Eurotia lanata</u>
Whitebark pine	<u>Pinus albicaulis</u>
Willow	<u>Salix</u> spp.
Yarrow	<u>Achillea millefolium</u>
Yellowbell	<u>Fritillaria pudica</u>

Appendix II

Average percent canopy cover and frequency of plants and bare ground area on a ridgeline elk concentration area and adjoining herbland slope.
(*, less than 1 percent)

Plants	Ridgetop		Ecotone ¹		Slope	
	Exclosure Plots:	17	YG1	YG3 25	YG4 25	YG5 25
Idaho fescue	13/76	15/53	11/80	1/28		
Slender wheatgrass	4/35	4/35	13/32	10/72	22/96	
Bluegrass	15/76	19/88	7/84	2/48	3/28	
Trisetum	1/18	1/23	1/40	3/56	1/16	
Needlegrass	2/12	*	*	*	*	
Oniongrass			2/12		3/40	
Alpine timothy		*	1/20	*		
Tufted hairgrass			3/16			
Mountain brome			*		4/36	
Sedge			12/68		*	
Other grass				1		
Grass totals	35	39	50	17	33	
Dandelion	68/100	69/100	40/100	15/20	2/36	
Cinquefoil	17/100	15/94	39/100	19/80	10/52	
Phlox	5/59	2/35	15/77	15/96		
Agoseris	3/47	1/24	2/28	4/52	3/56	
Yarrow	8/76	5/65	8/77	12/100	5/52	
Lupine	4/29	1/6	1/8		18/92	
Fleabane	4/29	*	10/56	3/52	3/44	
Elk thistle	2/59					
Parsley	1/6			*	*	
Hackelia	*	1/29			8/48	
Chickweed		2/29	1/28			
Plantain		*	4/40		*	
Buckwheat		*	1/20	1/12	1/12	
Astragalus			1/16			
Tall larkspur					4/16	
Other forbs	1	3				
Forb totals	113	99	122	69	54	
Plant totals	148	138	172	86	87	
Bare soil	12	10	26	61	51	

¹ Adjoining slope ecotone with edge of mountain meadow type.

Appendix III

Average percent canopy coverage and percent frequency of plant taxa on slope areas of the bunchgrass-shrub type (*, less than 1 percent).

No. Plots:	South					West ¹	SE ¹
	I 34	II 34	III 34	IV 34	V 34	VI 25	VII 25
Bluebunch wheatgrass	18/94	40/94	15/67	16/88	17/50	29/92	30/68
Bluegrass	6/56	15/88	5/44	1/12	1/6	18/88	14/32
Needle-and-thread	1/18	1/9					6/28
Ricegrass	6/47	*	9/47	3/35	4/18		
Junegrass		2/38	*			1/8	9/32
Slender wheatgrass			6/24				6/32
Brome	*			2/3			
Idaho fescue							13/32
Sedge		2/15					
Grass Totals	31	60	35	22	22	48	82
Astragalus	4/38		2/21	1/3		10/80	1/8
Bastard toadflax	1/15	*		*	2/29		
Fleabane	*	7/53		1/9			
Aster		3/35	4/38	1/9	5/38		
Elk Thistle		2/3					
Selaginella		3/9					
Hackelia				2/15			
Lupine				3/6			7/32
Balsamroot				2/6	2/9		
Buckwheat				1/9	1/12		
Flax					5/15		
Forb Totals	5	15	6	11	15	10	8
Phlox	6/74	5/65	6/50		1/6	6/84	
Leptodactylon	*	2/32		*			
Douglas rabbitbrush	1/15	3/29	2/32		1/3	23/92	
Rubber rabbitbrush	2/9	4/12	6/44		1/3		
Fringed sagewort	*	1/15	3/15				
Big sagebrush				3/9			22/64
Serviceberry				1/6	5/35		754 ²
Horsebrush					2/9		
Bitterbrush					*		17/36
Snowberry							7/36
Shrub Totals	9	15	17	4	10	29	46

¹ From Martinka (1965)

² Plants per acre

Appendix IV

Average percent canopy coverage and percent frequency of plant taxa on four 34-plot sample units within a 7,000 foot elevation forest park and a 9,000 foot subalpine meadow.
(*, less than 1 percent)

	Forest park			Subalpine
	Mesic center to Xeric edge			meadow
	I	II	III	IV
Rush	49/100			
Sedge	32/97	45/100	2/18	9/28
Tufted hairgrass	2/29	40/94	1/9	60/100
Bluegrass	1/12	*	*	*
Alpine timothy	*		5/76	*
Wild barley		2/21		
Slender wheatgrass		5/26	16/91	
Trisetum		1/12	*	
Idaho fescue			17/70	1/9
Mountain brome			14/94	
Oniongrass			1/29	
Grass Totals	84	93	56	70
Buttercup	1/15			
Gentian	2/47			
Lousewort	10/68			
Dandelion	6/70	5/24	34/100	28/100
Aster	16/94	20/73	4/47	1/21
Strawberry	3/24	1/12	3/36	
Valerian	2/15			
Parsley		2/21	*	
Yarrow		4/38	1/26	1/38
Cinquefoil		12/44	18/90	
Agoseris			2/50	
Horsetail			1/21	
Astragalus			9/38	
Buckwheat			1/24	*
Chickweed			1/12	
Collomia			1/21	
Flax			1/3	
Forb Totals	40	44	76	30
Willow	11/62			

Appendix V

Plants averaging 5 percent or more of the recorded use or used at 25 percent or more of the elk feeding sites on the bunchgrass-shrub type (percent/frequency).

Plants	Winter	Spring
Bluebunch wheatgrass	47/90	35/100
Needle-and-thread	11/65	5/58
Bluegrass	9/53	23/100
Ricegrass	3/33	4/54
Total grass	72	75
Balsamroot		8/29
Aster		2/41
Bastard toadflax		1/25
Total forbs	2	14
Douglas rabbitbrush	12/76	3/33
Rubber rabbitbrush	4/45	
Winterfat	2/33	
Fringed sagewort	1/43	
Leptodactylon		4/29
Phlox		2/33
Total shrubs	26	11

Appendix VI

Plants averaging 5 percent or more of the recorded use or used at 25 percent or more of the elk feeding sites on the valley meadow type (percent/frequency).

Plants	Winter	Spring	Summer	Fall
Sedge	34/75	18/95	6/44	1/33
Bluegrass	21/75	37/79		91/100
Rush	9/62			
Reedgrass	1/25			
Bentgrass			11/22	
Total grass	71	60	23	93
Dandelion		11/79	1/33	
Clover		7/53		
Agoseris		5/26		
Aster		4/68	1/56	4/66
Elk thistle				*/66
Total forbs	2	36	7	5
Willow	21/75	3/26	70/89	2/66
Shrubby cinquefoil	1/50			
Total shrubs	27	4	70	2

Appendix VII

Plants averaging 5 percent or more of the recorded use or used at 25 percent or more of the elk feeding sites on the deciduous and coniferous forest types (percent/frequency and * less than 5 percent).

	Deciduous forest			Coniferous forest			
	Winter	Spring	Summer	Winter	Spring	Summer	Fall
Bluegrass	33/75	24/56	2/59		1/28		
Needlegrass	3/25						*/27
Wheatgrass	1/75	5/56	5/65	9/56			
Brome	*/25	6/62	10/70	1/25			
Ricegrass	*/25						
Sedge	2/50	6/56	6/59	18/66	19/78	6/38	51/89
Rush	*/25						
Pinegrass		5/19	1/41	5/25	18/67	1/31	
Oniongrass			1/35				
Oatgrass				2/25			
Wild Rye				1/25			
Total Grass	43	50	25	37	40	10	52
Lupine	3/75	1/37	1/47		2/39	6/61	3/33
Aster	1/50	11/69	13/70	6/78	6/67	9/42	4/44
Goldenrod	*/25						
Sweetclover	1/50						
Horsetail	*/25						
Parsley		2/50					*/28
Clover		4/37					
Dandelion		5/62	6/59				11/46
Strawberry		1/31					
Geranium		8/50	2/59		3/44	2/35	
Agoseris		4/44	5/47		2/39	4/27	
Cinquefoil		1/25					
Buckwheat		1/25					
Microseris		1/31	1/29				
Yarrow			1/53				
Harebell			3/47				
Bedstraw			1/29				
Violet			1/25				
Balsamroot				1/44	2/33		
Arnica						11/44	
Fireweed							1/27
Total Forbs	6	49	50	9	39	47	8

Appendix VII (continued)

Willow	14/50		2/33				
Narrowleaf cottonwood	10/50						
Aspen	8/25	12/35	5/44				
Silverberry	5/75						
Serviceberry	4/25		4/55				
Bitterbrush	3/50						
Russet buffaloberry	2/50		6/55				
Silver sagebrush	2/25						
Kinnikinnick	1/50						
Rose	1/50	3/25	2/66				
Honeysuckle	1/50		2/25				
Snowberry		2/25	3/44				
Chokecherry			11/33				
Oregon grape			2/44				2/55
Lodgepole pine			2/33				
Limber pine			2/33				
Big sagebrush			1/25				
Douglas-fir			*/44				
Snowbush			5/28				8/22
Spirea			7/28	14/42			10/78
Huckleberry			5/10	12/50			5/44
Pachystima				5/19			13/44
Gooseberry				3/25			
Pipsissewa							1/33
Total Shrubs	51	1	25	54	21	43	40

Appendix VIII

Plants averaging 5 percent or more of the recorded use or used at 25 percent or more of the elk feeding sites on the sagebrush type (percent/frequency).

	Spring	Summer	Fall
Bluebunch Wheatgrass	9/65		20/70
Bluegrass	19/86		10/50
Idaho Fescue	11/56		4/30
Needlegrass	2/35		*/40
Brome	3/28		1/30
Junegrass	4/51		22/90
Sedge	4/51		
Oatgrass			10/30
Total Grass	54	4	74
Balsamroot	12/42	4/32	2/40
Buckwheat	4/46	6/47	1/40
Groundsel	4/35		
Lupine	3/25	50/68	18/70
Violet	2/46		
Microseris	2/60		
Agoseris	2/28	11/37	
Hawksbeard	1/42		
Parsley	1/30		
Little Sunflower		8/25	
Total Forbs	44	92	25
Total Shrubs	2	4	1

Appendix IX

Plants averaging 5 percent or more of the recorded use or used at 25 percent or more of the elk feeding sites on the herbland, subalpine meadow and forest park types (percent/frequency and *, less than 5 percent).

	Herbland			Subalpine Meadow		Burn		Forest Park		
	Spring	Summer	Fall	Summer	Summer	Spring	Summer	Spring	Summer	Fall
Slender Wheatgrass ¹	17/100	3/70	6/66		6/80	2/47		3/67	2/64	
Mountain Brome ²	19/80	5/60	34/100		25/100	6/73		9/76	2/33	1/33
Oniongrass	4/80	1/37			2/60			1/35		
Bluegrass	1/80	1/47	1/33	4/50		1/27		1/38	1/30	
Trisetum	1/60	1/43	1/33						1/36	
Sedge	5/40	2/50	3/33	33/78		20/100	1/66	11/81	5/64	55/100
Tufted Hairgrass				9/71						
Alpine Timothy				3/36				2/38	1/45	
Needlegrass					*/40	1/33				
Pinegrass					*/40					
Idaho Fescue								3/27		
Total Grass	48	15	45	50	57	13		33	14	56
Dandelion	1/40	13/73	3/33	7/50	2/80	17/93		8/49	24/82	
Aster	6/80	7/73	1/33	13/78	7/60	10/87		5/57	14/88	8/50
Agoseris	11/100	30/97			2/60	16/100		4/27	8/64	
Cinquefoil	7/60	10/70	7/33	2/36	2/40	9/87		5/54	9/82	
Violet	18/100	5/70			1/60	1/33		1/35		
Yarrow	*/60	1/40		1/36	1/40	1/27		*/30	1/25	
Parsley	1/60				*/40	*/27		1/62	1/36	
Buckwheat	1/40			1/43		1/27		3/30		
Yellowbell	1/40									
Hackelia	1/40	1/33	7/33							
Waterleaf	1/60									
Lupine		3/27	25/66	16/43	1/60	8/60			3/27	4/33
Tall Larkspur		9/53			1/40	7/47				
Geranium			2/33		13/100	5/60		10/46	3/39	
Plantain				4/43						

Appendix IX (continued)

Arnica					1/60	*/27			
Coneflower						1/27			
Astragalus						1/40		2/36	
Strawberry						1/60		1/30	
Little Sunflower							3/30		1/33
Balsamroot							8/25		
Elk Thistle								3/54	
Goldenrod									13/17
Total Forbs	52	85	55	50	42	83	66	82	29
Gooseberry						2/27			
Spirea									5/67
Oregon Grape									4/83
Pachystima									3/33
Aspen									*/33
Total Shrubs					1	4	1	4	15

¹ Includes other wheatgrass species on burn and forest park types.

² Includes other brome species on burn and forest park types.